

# Greenhouse Gas Mitigation Opportunities in California Agriculture

## Review of Emissions and Mitigation Potential of Animal Manure Management and Land Application of Manure

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## **ABSTRACT**

Manure management, primarily in anaerobic lagoons on dairies, is estimated to be the largest source of greenhouse gases from California agriculture. However, no field measurements from dairies in California have been published. A review of the broader literature revealed that emissions from anaerobic lagoons had more than 10 times the global warming potential of emissions from solid manure piles. Capping anaerobic lagoons and flaring the emitted methane, or fully converting to anaerobic digesters, could reduce total methane emissions by 92% (~7.7 Tg CO<sub>2</sub>e). Manure from farms is eventually applied to agricultural fields as fertilizer, where nitrous oxide is the primary greenhouse gas (GHG) emitted. Limited data are available on GHG emissions from manure-amended fields, and only two studies were conducted in California. This research suggested that fertilizing agricultural fields with manure rather than synthetic fertilizers results in lower GHG emissions as well as increased soil carbon storage. Despite the significance of dairy manure in GHG budgets at the state, regional, and global scale, this review reveals a surprising lack of field-scale research necessary to inform the development of best practices in California. Key areas of research for California include measurements of GHG emissions from dairy manure management systems and comparisons of GHG emissions from agricultural fields under different management practices.

## **Acknowledgments**

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## Contents

<b>ABSTRACT</b> .....	<b>2</b>
<b>INTRODUCTION</b> .....	<b>4</b>
<b>MANURE EMISSIONS</b> .....	<b>6</b>
<i>Background and Current Estimates</i> .....	6
<i>Approach to Data Compilation</i> .....	8
<i>Emissions from Manure Storage Systems</i> .....	10
<i>Emissions from Dairy Surfaces</i> .....	16
<i>Emissions from Whole Barns</i> .....	18
<i>Emissions from Whole Dairies</i> .....	19
<i>Implications for California Dairies</i> .....	20
Crusts .....	20
Anaerobic Digestion.....	21
Decreased Volume of Manure in Liquid Systems .....	21
<i>Life-cycle Assessments of Dairy Manure Management</i> .....	22
<b>LAND APPLICATION OF MANURE</b> .....	<b>23</b>
<i>Carbon Storage in Agricultural Fields through Manure Addition</i> .....	24
<i>Impacts of Manure Type</i> .....	25
<i>Limitations to Land Application of Manure</i> .....	26
<i>Application of Manure to Rangeland</i> .....	27
<i>Modeling Land Application of Manure</i> .....	27
<i>Implications for California Agriculture</i> .....	28
<b>KEY FINDINGS AND RESEARCH PRIORITIES</b> .....	<b>29</b>
<b>APPENDIX A: DATA FOR DAIRY EMISSIONS STUDIES</b> .....	<b>31</b>
<b>APPENDIX B: TERMINOLOGY</b> .....	<b>51</b>
<b>REFERENCES</b> .....	<b>52</b>

## INTRODUCTION

In California, greenhouse gas (GHG) emissions from agriculture account for approximately 7% of total GHG emissions from all sources in the state (CARB 2011). Emissions from livestock manure management are estimated to be the greatest contributor, accounting for one-third of agricultural emissions (Table 1; CARB 2011). California Air Resources Board (CARB) estimates indicate that dairy manure management is by far the largest source of GHG emissions from livestock manure management in California, accounting for over 97% of the methane (CH<sub>4</sub>) produced (Table 2; CARB 2011).

**Table 1. Greenhouse Gas Emissions from California Agriculture in Carbon Dioxide Equivalents (CO<sub>2</sub>e)**

Agricultural source	2009 emissions (Tg CO <sub>2</sub> e)
Manure management	10.34
Enteric fermentation	9.28
Soil management	9.02
Energy use	2.63
Rice cultivation	0.58
Histosol cultivation	0.16
Residue burning	0.06
<b>Total</b>	<b>32.07</b>

Source: CARB 2011.

Note: Emissions from manure spread on agricultural fields are included in soil management rather than manure management.

**Table 2. Estimated GHG Emissions for 2010 from Manure Management in Different Livestock Industries in California**

	Methane emissions (Tg CO <sub>2</sub> e)	Nitrous oxide emissions (Tg CO <sub>2</sub> e)
Dairy cattle	8.40037	1.12594
Beef cattle	0.02571	0.28108
Poultry	0.10508	0.04967
Swine	0.03588	0.00284
Goats	0.00008	0.00121
Horses	0.00398	0.02399
Sheep	0.00281	0.02354
<b>Total</b>	<b>8.57392</b>	<b>1.50826</b>

Source: CARB 2011.

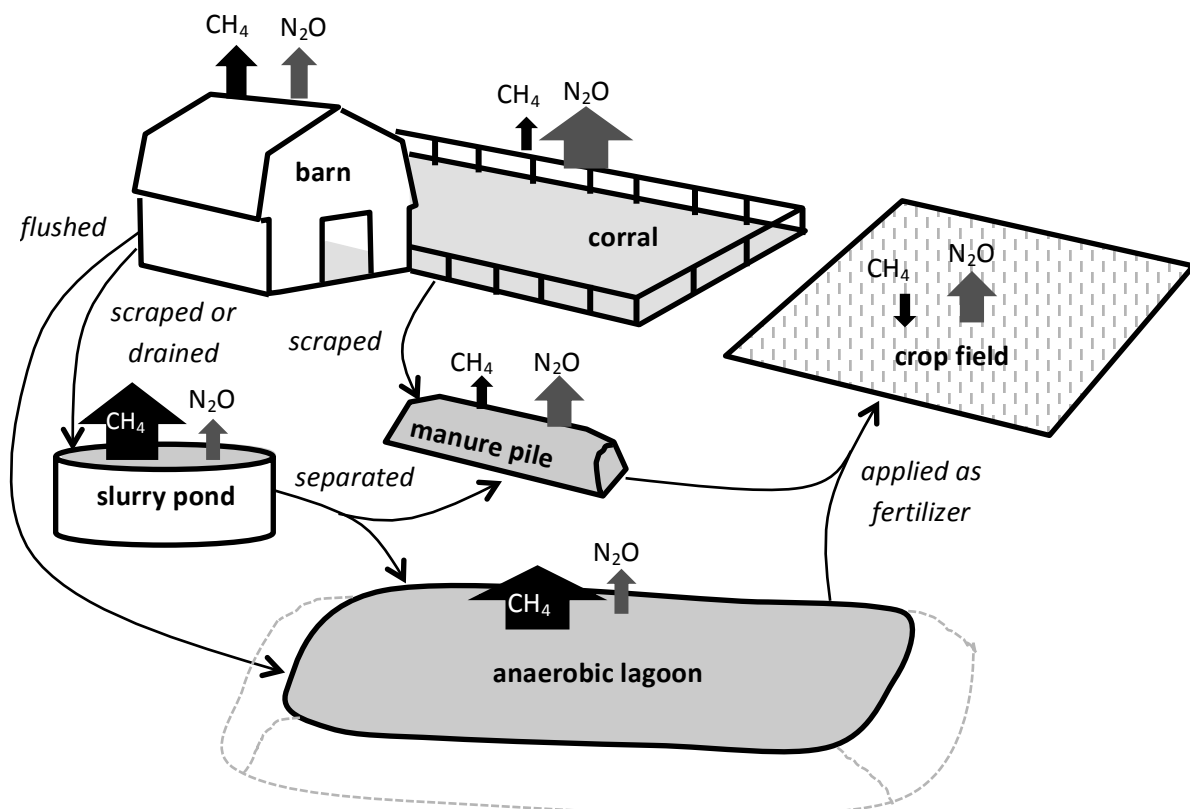
Note: Manure management does not include manure deposited or applied to pasture or agricultural fields.

Many areas on dairies are potential sources of greenhouse gases (Figure 1, USEPA 2011a; Meyer et al. 2011). The abundant carbon (C), nitrogen (N), and water in liquid and solid livestock waste promote microbial activity that drives the production of carbon dioxide (CO<sub>2</sub>), CH<sub>4</sub>, and nitrous oxide (N<sub>2</sub>O). Carbon dioxide is primarily produced through aerobic microbial respiration associated with decomposition. Most C accounting schemes use aerobic decomposition as a baseline, thus C losses through aerobic respiration as CO<sub>2</sub> are not counted as part of the global warming potential. Methane is produced when the organic matter (including manure) is decomposed anaerobically by methanogens. Net CH<sub>4</sub> emissions are promoted under warmer temperatures, high N availability, and low oxygen

availability—conditions typical of manure storage areas. Methane is oxidized by aerobic and anaerobic methanotrophs. Methane consumption is common at anaerobic-aerobic interfaces, and upland soils are generally a net sink for CH<sub>4</sub> (Conrad 2007).

When organic matter is decomposed, organic N is mineralized into plant available forms such as ammonium (NH<sub>4</sub><sup>+</sup>, a potential air pollution source when it volatilizes as NH<sub>3</sub> and nitrate (NO<sub>3</sub><sup>-</sup>, a potential water pollutant when it leaches out of the soil). Nitrous oxide production occurs as an intermediate step during two microbial processes: nitrification, the aerobic oxidation of NH<sub>4</sub><sup>+</sup> to NO<sub>3</sub><sup>-</sup>, and denitrification, the anaerobic reduction of NO<sub>3</sub><sup>-</sup> to dinitrogen gas (N<sub>2</sub>, the inert gas comprising the bulk of Earth’s atmosphere; Firestone and Davidson 1989). Nitrous oxide production is greatest when aerobic and anaerobic conditions occur in close proximity, as in manure piles and corral soils.

**Figure 1. Sources of N<sub>2</sub>O and CH<sub>4</sub> on Dairies**



*Note:* Thin arrows indicate movement of manure between locations. Block arrows indicate CH<sub>4</sub> and N<sub>2</sub>O fluxes; thicker arrows indicate relatively greater emission rates.

Land application of manure directly from cows onto pasture or rangeland, or application to crop fields following on-dairy management, can be a significant source of greenhouse gases but may also contribute to soil C sequestration (DeLonge et al. 2013). Land-applied manure in California contained an estimated 231 Gg (1 Gg = 10<sup>9</sup> g) of N in 2009 and resulted in direct N<sub>2</sub>O emissions of ~4 Gg, using the IPCC (2006) emissions factor of 0.01 g N<sub>2</sub>O-N g manure-N<sup>-1</sup>, equivalent to 1.083 Tg CO<sub>2</sub>e (1 Tg = 10<sup>12</sup> g; CARB 2011). Methane fluxes on agricultural fields are not considered by regional and countrywide GHG accounting methods except for rice cultivation (IPCC 2006; USEPA 2011a), so no estimates are available for the statewide CH<sub>4</sub> emissions associated with land application of manure in California.



Reductions in GHG emissions from California agriculture, particularly dairy manure management, can contribute to the state's climate change mitigation goals. Following the approach of the IPCC, this report considers separately the GHG emissions associated with on-farm manure management and those associated with manure application to agricultural fields and rangelands. For each, this report summarizes current estimates and field measurements, identifies the greatest sources of greenhouse gases, and evaluates the biophysical mitigation potential of alternative management practices.

## **MANURE EMISSIONS**

### ***Background and Current Estimates***

Dairies' manure management depends on whether manure is handled as solid or liquid, where it was deposited on the farm, how it will be used, and regional preferences for storage. Cows deposit manure onto dairy surfaces in dry lots, hardstandings, barn floors, and loafing pens. The length of time the cows are present and the frequency with which the manure is removed (by scraping, flushing, or vacuuming) are important determinants of the GHG fluxes. Liquid manure is generally stored in lagoons, slurry tanks, or settling ponds. Lagoons are not stirred, and anaerobic conditions develop rapidly. Slurry tanks have a larger solids content than lagoons; some solids typically float on the surface and form a surface crust. The crust provides a substrate that spans anaerobic and aerobic environments wherein N<sub>2</sub>O production and CH<sub>4</sub> oxidation can both occur (Petersen, Amon, and Gattinger 2005; Petersen and Sommer 2011). Solid manure is composed of the solids scraped from dairy surfaces, raked from loafing pens, or separated from slurry. It is generally stored in outdoor, uncovered piles, sometimes inside dry lot corrals. Solid manure piles are heterogeneous in composition and can contain both aerobic and anaerobic areas, depending on moisture content and management practices. Loafing pens are typically smaller sources of manure, because milk cows are in them only when they are not in milking parlors or barn stalls.

Publicly available information about dairy manure management practices in California is limited. The USEPA (2011a) estimated GHG emissions from dairies using data on manure management practices calculated in ERG (2000) and adjusting for changes in the dairy industry (growth and consolidation) over time (Table 3). The 2009 estimates by CARB (2011) of the proportions of manure in each management system are similar to those of the USEPA (Table 3). The estimated distribution of manure management in California was typical of the distribution in the southwest United States, which relies heavily on anaerobic lagoons. Survey results of Meyer et al. (2011) are comparable, though they suggest the use of daily spread may be overestimated. Because Meyer et al. (2011) did not use the same categories as the EPA, the following conversions were used:

- The percent of dairies using solid storage was estimated as 100 minus the percent of dairies that did not separate solid and liquid manure.
- Liquid slurry use was calculated as equal to the percent of dairies that didn't separate solid and liquid manure.
- Anaerobic lagoon use was calculated as equal to the percent of dairies with storage ponds minus the liquid/slurry percent.
- Dry lot was estimated as equal to the percent of dairies where manure solids were piled.

The allocation between solid and liquid storage systems varies among dairies depending on infrastructure and management preferences and within a dairy depending on seasonal evaporation losses and management decisions (UC 2005).

**Table 3. Manure Management Distribution (%) for Dairies and Heifer Facilities (Farms with No Milk Cows or Milking Facilities) in California and Glen and Tulare Counties**

	Dairies					Dairy heifer facilities			
	Pasture	Daily spread	Solid storage	Liquid/slurry	Anaerobic lagoon	Daily spread	Dry lot	Liquid/slurry	Pasture
CA (USEPA)	1	11	9	20	59	11	88	1	1
CA (CARB)	0.93	10.8	9.24	20.1	56.5	10.8	88.3	0.87	0.93
Glen County	Na	0	63.2	36.8	52.7	0	80.1	na	na
Tulare County	Na	0	70.5	29.5	66.0	0	93.2	na	na

Sources: Glen and Tulare Counties: Meyer et al. (2011), California: USEPA (2011a) and CARB (2011).

Note: The USEPA (2011a) and CARB (2011) accountings were forced to add up to 100, whereas the survey by Meyer et al. (2011) allowed farmers to report all practices used and do not add up to 100. na = not available.

Based on the allocations in Table 3, CARB (2011) estimated GHG emissions by manure management system for dairy cows and dairy heifers using a Tier 2 approach (Table 4). The results reveal that the overwhelming majority of dairy cow manure GHG emissions are derived from liquid manure management and that most of these emissions are CH<sub>4</sub>.

**Table 4. Estimates of the Number of Cows Contributing Manure to Each Management System and the CH<sub>4</sub> and N<sub>2</sub>O Emissions from That Manure Management in Tg CO<sub>2</sub>e for 2009**

Management practice	Head	CH <sub>4</sub> (Tg CO <sub>2</sub> e)	N <sub>2</sub> O (Tg CO <sub>2</sub> e)	Emissions per cow (g CO <sub>2</sub> e/hd/d)
<b>Dairy cows</b>				
Anaerobic lagoon	1,039,082	7.323	0.348	20,227
Liquid/slurry	370,093	1.181	0.212	10,311
Deep pit	4,186	0.006	0.001	4,370
Anaerobic digester	40,786	0.033	0.007	2,685
Solid storage	170,073	0.061	0.096	2,528
Daily spread	198,755	0.009	0.014	321
Pasture	17,024	0.002	0	271
<b>Dairy heifers</b>				
Liquid/slurry	6,843	0.007	0.002	3,538
Dry lot	691,174	0.030	0.497	2,090
Daily spread	84,570	0.001	0.003	132
Pasture	7,244	0.000	0	122
<i>Total dairy cows</i>	<i>1,840,000</i>	<i>8.615</i>	<i>0.678</i>	<i>40,712</i>
<i>Total dairy heifers</i>	<i>782,920</i>	<i>0.039</i>	<i>0.502</i>	<i>5,882</i>

Source: CARB (2011).

Note: Nitrous oxide emissions from manure directly deposited on pasture are counted in "Agricultural Soil Management" and therefore are reported as 0 here.

Though useful, the data presented in Table 4 reflect two major uncertainties. First, manure is often managed in multiple ways for one cow (Figure 1; Meyer et al. 2011); how, if at all, CARB (2011) accounts for this fact is unclear. Second, the emissions estimates calculated by Tier 2 approaches are based on equations and emissions factors derived largely from laboratory incubations or pilot-scale studies conducted under a limited range of conditions. Thus, their applicability to farm-scale operations under a range of environmental conditions has not been confirmed. This research cannot address the first issue, but it explores the second by compiling field-scale emissions data and comparing it to Tier 2 calculations.

### ***Approach to Data Compilation***

Published field measurements of GHG emissions from on-dairy manure management on working and research dairies globally were reviewed and compared with emissions rates calculated using approaches and regional estimates from the IPCC (2006) and USEPA (2011a). Three broad categories of measurement areas were considered: (1) manure management (including liquid and solid systems), (2) holding area surfaces (dry lot corrals, loafing pens, hardstandings, and barn floors), and (3) whole barns. Following the approach of the IPCC (2006), liquid manure emissions data were divided into two groups: (1) anaerobic lagoons and (2) slurry tanks and settling ponds. Data from pilot and lab-scale studies were excluded. Also excluded were data from studies on beef cattle facilities, because their management and herd characteristics differ from those of dairies. Greenhouse gas emissions from anaerobic digesters are particularly difficult to quantify due to the variety of systems and operation practices (ex. retention times, effectiveness of seals in preventing gas escape, and composition of material entering the digester; Massé et al. 2011). Thus, the literature related to anaerobic digester emissions is not reviewed.

Thirty-two studies met the above criteria (Table 5). Most of these studies were conducted in North America and Europe (see Appendix Table 1A for data on location, measurement technique, and climate). No data were available for California, although results of at least two studies of GHG emissions from dairy sources in California are not yet published. Greenhouse gases were measured from the California dairy barn monitored for the National Air Emissions Monitoring Study (NAEMS) project (Zhao et al. 2010; F. Mitloehner, personal communication). Additionally, an on-farm study is measuring GHG emissions from a range of dairy sources to determine mechanisms controlling fluxes (Owen and Silver).

Methane flux measurements that included enteric fermentation by cows in barns, corrals, or whole dairies were corrected by subtracting  $100 \text{ kg CH}_4 \text{ hd}^{-1} \text{ y}^{-1}$  (hd = head) (Mitloehner, Sun, and Karlik 2009), a moderate estimate compared with other estimates (Kinsman et al. 1995; IPCC 2006). Enteric emissions are dependent on many highly variable cow and feed characteristics, but the uncertainty in enteric emissions rates is smaller than that in the emission rates from manure management, as shown below.

Most studies included measurements of one source at different times (e.g., seasonally), or measurements from areas in which some management component was different (e.g., mixed versus static manure piles, barns with different flooring and scraping mechanisms), or both. Each permutation was included in the compilation. The average emissions rate for a given dairy source was calculated first by averaging the emissions rates compiled from each paper (though separate sources measured in the same study, such as primary and secondary lagoons, were treated separately and not averaged together) and then by averaging those values. This method avoided weighting the value toward studies and management practices with comparatively more measurements.

Emissions rates are presented in the units of  $\text{g trace gas hd}^{-1} \text{ d}^{-1}$  and  $\text{kg trace gas ha}^{-1} \text{ d}^{-1}$  when possible. Because data were not always provided to enable conversions between units, the average emissions rates presented below may include different studies depending on the units in which the emissions could be

reported. Data are reported as g CH<sub>4</sub> and g N<sub>2</sub>O rather than g CH<sub>4</sub>-C and g N<sub>2</sub>O-N. Some have argued that expressing emissions per kg of fat-corrected milk is a better way to capture the efficiency of dairy production (Place and Mitloehner 2010; O'Brien et al. 2012), but few studies provided the information necessary to calculate these values. Similarly, livestock or animal units, which are the number of animals normalized by the typical mass of the animal (500 or 600 kg is usually used for dairy cows), are sometimes used. Given the lack of consistency in conversions and lack of information for some studies, livestock units were converted to head count when necessary. Methane and N<sub>2</sub>O have different global warming potentials once in the atmosphere. To compare the total global warming potential of the GHG emissions from the measured areas, N<sub>2</sub>O and CH<sub>4</sub> emissions rates were converted to 100-year CO<sub>2</sub>e by multiplying by 298 and 25, respectively (IPCC 2006) and summing the two.

**Table 5. Studies Reporting GHG Emissions Measurements from Working or Research Dairies and the Areas Measured by Each**

Study	Lagoon	Slurry tank	Manure pile	Compost area	Open lot	Concrete pens	Barn	Whole dairy
Borhan et al. (2011a)	x			x	x		x	
Leytem et al. (2011)	x			x	x			
Borhan et al. (2011b)	x				x		x	
Bjorneberg et al. (2009)	x				x			
Craggs et al. (2008)	x							
Safley and Westerman (1988)	x							
Safley and Westerman (1992)	x							
Todd et al. (2011)	x							
Husted (1994)		x	x					
Sneath et al. (2006)		x	x					
Hensen et al. (2006)		x						x
Kaharabata et al. (1998)		x						
VanderZaag et al. (2011)		x						
Ahn et al. (2011)			x					
Amon et al. (2006)			x					
Brown et al. (2002)			x					
Gupta et al. (2007)			x					
Osada et al. (2001)			x					
Sommer et al. (2004)			x					
Kaharabata et al. (2000)					x			
Ellis et al. (2001)						x		
Gao et al. (2011)						x		
Misselbrook et al. (2001)						x		
Adviento-Borbe et al. (2010)							x	
Kinsman et al. (1995)							x	
Marik and Levin (1996)							x	
Ngwabie et al. (2009)							x	
Samer et al. (2012)							x	

Snell et al. (2003)	x
van Vliet et al. (2004)	x
Zhang et al. (2005)	x
McGinn and Beauchemin (2012)	x

The IPCC (2006) Tier 2 approach, used with values from the IPCC (2006) and the USEPA (2011a) and equations 1 and 2 below, was applied to predict CH<sub>4</sub> and N<sub>2</sub>O emissions rates for each source. The resulting values were then compared with the average field measurements. The IPCC Tier 2 method calculates CH<sub>4</sub> emissions (EF, g CH<sub>4</sub> hd<sup>-1</sup> y<sup>-1</sup>) on the basis of the volatile solid content of manure excreted by cows (VS, kg VS hd<sup>-1</sup> y<sup>-1</sup>), a CH<sub>4</sub> conversion factor (MCF, %) based on the manure management practice, and the maximum possible CH<sub>4</sub> production rate from the volatile solids in the manure (B<sub>o</sub>, m<sup>3</sup> CH<sub>4</sub> kg VS<sup>-1</sup>)(IPCC 2006):

$$EF_{CH_4} = VS \times MCF/100 \times B_o \times 662 \text{ g CH}_4 \text{ (m}^3 \text{ CH}_4\text{)}^{-1} \quad (\text{eq. 1})$$

The IPCC Tier 2 method calculates direct N<sub>2</sub>O emissions on the basis of an emissions factor ( $EF_{N_2O}$ , kg N<sub>2</sub>O-N kg N excreted<sup>-1</sup>) and annual N excretion rates, which are a function of cows' energy intake, the percent of crude protein, the milk production rate and protein content, cow growth, typical animal mass, and (IPCC equations 10.31, 10.32, and 10.33; IPCC 2006).  $EF_{N_2O}$  can be converted into N<sub>2</sub>O emissions rates equivalent to those measured here ( $N_{2O_D}$ , g N<sub>2</sub>O hd<sup>-1</sup> d<sup>-1</sup>, where the subscript D refers to direct emissions) using the typical animal mass ( $TAM$ , kg) and country- or region-specific N excretion rates ( $N_{ex}$ , kg N 1000 kg  $TAM^{-1}$  d<sup>-1</sup>):

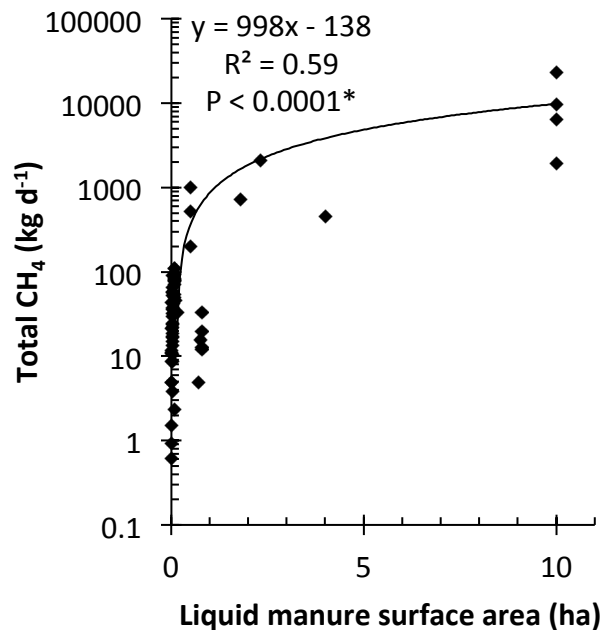
$$N_{2O_D} = EF_{N_2O} \times 44/28 \times TAM \times N_{ex} \quad (\text{eq. 2}).$$

Indirect N<sub>2</sub>O emissions, derived from the oxidation of gaseous emissions such as ammonia (NH<sub>3</sub>) and nitrous oxides (NO<sub>x</sub>), are important for calculating the amount of N remaining in manure for its use as an organic fertilizer (IPCC 2006). No indirect N<sub>2</sub>O emissions are included in the calculations here; values reported here should be considered minimum values. The complete data compilation is available in appendix tables A2–8.

### **Emissions from Manure Storage Systems**

Consistent with IPCC and EPA estimates, the greatest emissions on dairies were measured from manure storage. Total emissions from liquid systems averaged ~19,000 g CO<sub>2</sub>e hd<sup>-1</sup> d<sup>-1</sup>, and those from solid systems averaged ~1,700 g CO<sub>2</sub>e hd<sup>-1</sup> d<sup>-1</sup> (Table 6). Liquid manure systems had greater CH<sub>4</sub> emissions than solid systems and smaller, but non-zero, N<sub>2</sub>O emissions (Table 6). By contrast, IPCC emissions factors predicted zero N<sub>2</sub>O emissions from liquid manure except for slurry systems with a crust (IPCC 2006). The range of GHG emissions from anaerobic lagoons and slurry tanks were similar, although anaerobic lagoons tended to emit more CH<sub>4</sub> on a per head basis than slurry tanks (Table 6). Total CH<sub>4</sub> emissions (kg d<sup>-1</sup>) from liquid manure systems were positively correlated with surface area ( $r^2 = 0.59$ ,  $P < 0.0001$ , Figure 2). Despite the importance of manure composition in emissions calculations, few studies reported volatile solid content or other measures (tables A2 and A3).

**Figure 2. Total Daily CH<sub>4</sub> Emissions from Liquid Manure Management Increased with Surface Area**



Tier 2-predicted CH<sub>4</sub> emissions from anaerobic lagoons were 892 g hd<sup>-1</sup> d<sup>-1</sup>, < 20% below and within the standard error of the field measurement mean of 1097 ± 591 g hd<sup>-1</sup> d<sup>-1</sup> (Table 7). Slurry tank and pond CH<sub>4</sub> emissions were underestimated by the predicted rate of 81 g hd<sup>-1</sup> d<sup>-1</sup>, which is less than a third of the field mean of 276 ± 129 g hd<sup>-1</sup> d<sup>-1</sup> (Table 7). The predicted CH<sub>4</sub> emissions rates may not adequately account for the emissions related to field-scale events such as periodic tank emptying and surface crust disturbances. Crust disturbances by rainfall increased CH<sub>4</sub> emissions by 1.2–4 times in one study (Kaharabata et al. 1998), whereas crust thawing increased CH<sub>4</sub> emissions over an order of magnitude (VanderZaag et al. 2011). These episodic releases of CH<sub>4</sub> comprise a large portion of the total CH<sub>4</sub> emissions but are not easily included in emission factors based on environmental characteristics or region. Additionally, slurry remaining in the tank after emptying inoculates fresh slurry with microbes, generally increasing CH<sub>4</sub> emissions and decreasing the time to the onset of CH<sub>4</sub> production once the tank is refilled (Zeeman 1991), although this effect may be less important in cold climates (VanderZaag et al. 2011). Volatile solid carryover is considered in the USEPA default emissions factors (Mangino et al. 2002; USEPA 2011a) but not in the IPCC emissions factors.

Non-zero, direct N<sub>2</sub>O emissions were always detected from anaerobic lagoons, with a mean of 1.4 ± 1.2 g hd<sup>-1</sup> d<sup>-1</sup> (tables 6 and A2), and thus the IPCC N<sub>2</sub>O emissions factor of zero was an underestimate (Table 8). Scaling these emissions to anaerobic lagoon use in California in 2009 (Table 4) suggests *additional* N<sub>2</sub>O emissions from anaerobic lagoons of 0.16±0.14 Tg CO<sub>2</sub>e, or a mean increase of nearly 50% over the CARB (2011) estimate. Tier 2-predicted slurry tank N<sub>2</sub>O emissions of 2.1 g hd<sup>-1</sup> d<sup>-1</sup> were similar to the field mean of 0.7 ± 0.7 g hd<sup>-1</sup> d<sup>-1</sup> when the emissions factor for slurry with a natural crust was used, as was the case for most of the slurry tanks (tables 8 and A3).

The comparison between the Tier 2 values and the field measurement means suggests that anaerobic lagoon emissions are reasonably well predicted and that estimates for California are likely to be reliable

(tables 7 and 8). Slurry tank emissions estimates appear less reliable and suggest the California value may be an underestimate.

Manure pile emissions were measured in nine studies, and a large manure composting area was measured in one (tables A1 and A4). Few studies reported manure characteristics, but of those that did, the dry matter (DM) content was 19–45%, manure N content was 0.11–3.92 g N kg DM<sup>-1</sup>, manure C content was 4.8–65 g C kg DM<sup>-1</sup>, and the volatile solid content of manure was 16–93% (Table A4). Although the sample size was low, these ranges reveal the variety and heterogeneity in chemical and physical characteristics of solid manure. Data from Gupta et al. (2007) were excluded from Table 6 because they were reported in units of mass trace gas per mass manure per time, without adequate information to confidently convert to the units used here. The estimates of mass gas hd<sup>-1</sup> d<sup>-1</sup> in Table A4 were calculated assuming 2.6 kg DM manure production hd<sup>-1</sup> d<sup>-1</sup>, the IPCC (2006) default for volatile solid production by dairy cows in India.

Greenhouse gas emissions from manure piles spanned 3 to 4 orders of magnitude (Table 6), and much of this variation appeared to be related to measurement technique rather than variation among piles. Studies that used sample masts, tracer gas techniques, or large dynamic flux chambers measured higher emissions than studies using smaller dynamic or static chambers (Table A4). For example, CH<sub>4</sub> emissions measured using large footprint techniques were 93 ± 25 kg ha<sup>-1</sup> d<sup>-1</sup> (n = 5), or 104 g hd<sup>-1</sup> d<sup>-1</sup> (n = 1), compared with 0.091 kg ha<sup>-1</sup> d<sup>-1</sup> (n = 1), or 0.023 ± 0.012 g hd<sup>-1</sup> d<sup>-1</sup> (n = 3), from small chambers. The number of studies is extremely low, however, and more data are needed to confirm this observation. Theoretically, measurement methods that measure gas fluxes from larger areas likely have greater capacity to integrate the spatial and temporal variability of pile emissions arising from manure piles' varied composition and oxygen availability (Parkin and Kaspar 2004; Sommer et al. 2004).

Methane emissions were greatest when manure piles were first constructed. Data were inadequate to determine if subsequent additions to the piles also resulted in greater emissions. Mixed piles had greater CH<sub>4</sub> emissions than static ones, consistent with the findings of pilot-scale studies (Yamulki 2006), but contrary to expectations that mixing would aerate the pile and decrease CH<sub>4</sub> production. The predicted CH<sub>4</sub> emissions rate from piles was 6.0 g hd<sup>-1</sup> d<sup>-1</sup> (Table 7), which greatly underestimated the large-footprint measurements and overestimated the small chamber measurements. Thus, the California 2009 estimate (Table 7) may also be an underestimate.

Manure piles had high N<sub>2</sub>O emissions, with a field measurement mean second only to corrals (Table 6). Ahn et al. (2011) found that N<sub>2</sub>O emissions increased when the piles were mixed. The predicted N<sub>2</sub>O emissions rate was 1.2 g hd<sup>-1</sup> d<sup>-1</sup>, within the standard error of the field measurement mean (Table 8). The California 2009 estimate was higher, partially because it includes indirect N<sub>2</sub>O emissions, which arise from the deposition of manure-derived NH<sub>3</sub> on soil, where it is converted to N<sub>2</sub>O. Increasing pile aeration may decrease N<sub>2</sub>O emissions, but more data are needed to confirm this hypothesis.

**Table 6. Summary of the Means and Ranges of N<sub>2</sub>O and CH<sub>4</sub> Emissions Rates Measured by the Studies in Table 1**

	Mean ± standard error (n) Range						Global warming potential = 25 x CH <sub>4</sub> + 298 x N <sub>2</sub> O Mean ± standard error	
	(kg/ha/d)			(g/ha/d)			CO <sub>2</sub> e (kg/ha/d)	CO <sub>2</sub> e (g/ha/d)
	CO <sub>2</sub>	N <sub>2</sub> O	CH <sub>4</sub>	CO <sub>2</sub>	N <sub>2</sub> O	CH <sub>4</sub>		
<b>Manure management*</b>	3987 ± 1349 (12) 52 - 126308	4.5 ± 1.5 (13) 0.00 - 27	485 ± 116 (20) 0 - 2307	1929 ± 576 (10) 13 - 6575	1.7 ± 0.7 (10) 0 - 8322	619 ± 286 (18) 0 - 7710	13465 ± 3344	15985 ± 7380
<b>Solid manure piles*</b>	7334 ± 2816 (4) 367 - 12631	7.0 ± 2.4 (7) 0.13 - 27	78 ± 25 (6) 0.1 - 352	2066 ± 1904 (2) 162 - 9716	2.9 ± 1.8 (4) 0.06 - 21	35 ± 30 (4) 0 - 271	4039 ± 1332	1732 ± 1287
<b>Liquid</b>	1755 ± 943 (8) 52 - 8548	1.5 ± 0.9 (6) 0.027 - 10.3	660 ± 142 (14) 0 - 2307	1883 ± 728 (8) 13 - 6575	1.1 ± 0.6 (6) 0.010 - 6.8	745 ± 350 (14) 0 - 7710	16944 ± 3803	18935 ± 8952
<b>Slurry tanks and ponds</b>	nm	1.3 ± 1.1 (3) 0 - 19	756 ± 268 (6) 0 - 2027	nm	0.7 ± 0.7 (3) 0 - 12.3	276 ± 129 (6) 0 - 899	19280 ± 7045	7117 ± 3442
<b>Anaerobic lagoons</b>	1755 ± 943 (6) 52 - 8548	1.7 ± 1.6 (3) 0.027 - 8.5	588 ± 158 (8) 7.0 - 2307	1883 ± 728 (6) 13 - 6575	1.4 ± 1.2 (3) 0.010 - 6.5	1097 ± 591 (8) 10 - 7710	15207 ± 4421	27822 ± 15125
<b>Surfaces</b>	1349 ± 637 (7) 137 - 10007	0.50 ± 0.21 (9) 0.002 - 5.98	2.81 ± 9.47 (12) -90.1 - 118	5138 ± 3895 (7) 69 - 55594	1.87 ± 1.17 (9) 0 - 33.2	5.78 ± 26.9 (12) -222 - 653	215 ± 299	703 ± 1022
<b>Corrals**</b>	2263 ± 1411 (3) 324 - 10007	1.18 ± 0.35 (3) 0.033 - 5.98	-9.5 ± 21 (5) -90.1 - 118	11623 ± 8329 (3) 366 - 55594	5.5 ± 2.5 (3) 0.2 - 33	0.4 ± 68 (5) -222 - 653	115 ± 639	1658 ± 2464
<b>Hard-standings**</b>	nm	0.009 ± 0.002 (2) 0.002 - 0.030	20 ± 17 (3) 0.05 - 59	nm	0.0012 ± 0.0002 (2) 0.0002 - 0.003	19 ± 2.1 (3) 0.01 - 67	493 ± 428	481 ± 54
<b>Barn floor</b>	583 ± 263 (4) 137 - 1596	0.23 ± 0.16 (4) 0.01 - 1.0	5.6 ± 4.3 (4) 0.4 - 28	257 ± 108 (4) 69 - 684	0.07 ± 0.04 (4) 0.004 - 0.3	2.4 ± 1.8 (4) 0.1 - 12	209 ± 153	81 ± 56
<b>Whole barn**</b>	21206 ± 15824 (3) 1060 - 101738	33 ± 30 (2) 0 - 63	158 ± 80 (7) -80 - 887	19737 ± 15087 (3) 747 - 96048	31 ± 29 (2) 0 - 61	153 ± 75 (7) -70 - 837	13657 ± 10995	13170 ± 10524
<b>Whole dairy**</b>	nm	nm	nm	nm	nm	309 ± 96 (18) -203 - 1005	nm	7735 ± 2405

\*Excludes data from Gupta et al. (2007), which could not with confidence be converted into the units used here.

\*\*Where necessary, CH<sub>4</sub> emissions rates have been corrected for enteric emissions assuming 100 kg hd<sup>-1</sup> y<sup>-1</sup>. Carbon dioxide emissions were not corrected for respiration.

*Note:* The mean for a measurement area was calculated by first averaging measurements within studies and then averaging across studies (n = number of studies). The exception was whole dairy measurements, which were treated individually and not grouped by study (i.e., n = number of dairies measured). In this table, values for the manure management row are the average of solid and liquid manure management systems, values for the liquid row are the average of all liquid systems (slurry tanks, ponds, and anaerobic lagoons), and values for the surfaces row are the average of corrals, hard-standings, and barn floors. nm = not measured.



**Table 7. Calculation of Predicted CH<sub>4</sub> Emissions Rates Using the Tier 2 Approach (eq. 1) and Comparison with Field-Measured Values (Mean ± Standard Error (n)) and the 2009 Estimates for California from CARB (2011)**

	VS (kg/hd /y)	VS source	MCF (%)	MCF source	B <sub>0</sub> (m <sup>3</sup> /k g)	B <sub>0</sub> source	CH <sub>4</sub> emissions rate (g/hd/d)		
							Tier 2 predicted	Field measured	California estimate
<b>Anaerobic lagoons</b>	2770	USEPA (2011a), average ID, NM, TX, NC	74	USEPA (2011a), average ID, NM, TX, NC	0.24	IPCC (2006)/USEPA (2011a)	892	1097 ± 591 (8)	772
<b>Slurry tanks and ponds</b>	1861.5	IPCC (2006), Western Europe	10	IPCC (2006), MAT<10, with crust	0.24	IPCC (2006), Western Europe	81	276 ± 129 (6)	350
<b>Manure pile</b>	2750	USEPA (2011a), average TX and ID	0.5	USEPA (2011)	0.24	USEPA (2011a)	6.0	35 ± 30 (4)	39
<b>Corral</b>	2800	USEPA (2011a)	1	IPCC (2006)/USEPA (2011a), cool MAT	0.24	IPCC (2006)/USEPA (2011a)	12.2	0.4 ± 68 (5)	4.8
<b>Concrete hardstanding</b>	1861.5	IPCC (2006), Western Europe	1	IPCC (2006), cool MAT	0.24	IPCC (2006), Western Europe	8.1	3 ± 3 (2)	NA
<b>Brick hardstanding</b>	1022	IPCC (2006), Asia	1	IPCC (2006), cool MAT	0.13	IPCC (2006), Asia	2.4	52 ± 21 (1)	NA
<b>Whole barn</b>	1861.5	IPCC (2006), Western Europe	3	IPCC (2006), cool MAT	0.24	IPCC (2006), Western Europe	24	153 ± 75 (7)	158
<b>Whole dairy (= corral + manure pile + slurry tank + whole barn)</b>							117	309 ± 96 (18)	816

Note: VS = volatile solids, MCF = methane conversion factor, B<sub>0</sub> = the maximum possible CH<sub>4</sub> production rate from the volatile solids in the manure, NA = not available. The 2009 California deep pit estimate was used as the whole barn value, and the whole dairy value was the sum of anaerobic lagoons, corrals, and manure piles.

**Table 8. Calculation of N<sub>2</sub>O Emissions Rates Using the Tier 2 Approach (Eq. 2) and Comparison with Field-Measured Values (Mean ± Standard Error (n)) and the 2009 Estimates for California from CARB (2011)**

	EF <sub>N<sub>2</sub>O</sub> (kg N <sub>2</sub> O-N/ kg N excreted)	TAM (kg/hd)	N <sub>ex</sub> (kg N/ 1000kg TAM/d)	EF <sub>N<sub>2</sub>O</sub> and N <sub>ex</sub> sources	N <sub>2</sub> O emissions rate (g/hd/d)		
					Tier 2 predicted	Field measured	California estimate
<b>Anaerobic lagoon</b>	0	600	0.25	USEPA (2011a), average of ID and TX	0	1.4 ± 1.2 (3)	3.1
<b>Slurry tanks and ponds</b>	0.005	600	0.44	IPCC (2006), North America, with crust	2.1	0.7 ± 0.7 (3)	5.3
<b>Solid manure pile</b>	0.005	600	0.26	USEPA (2011a), ID	1.2	2.9 ± 1.8 (4)	5.2
<b>Corral</b>	0.02	600	0.25	USEPA (2011a), average of ID and TX	4.8	5.5 ± 2.5 (3)	6.6
<b>Concrete hardstanding</b>	0.02	600	0.44	IPCC (2006)	8.3	0.0012 ± 0.0002 (2)	NA
<b>Whole barn</b>	0.002	600	0.48	IPCC (2006), western Europe	0.91	31 ± 29 (2)	1.4

*Note:* EF<sub>N<sub>2</sub>O</sub> = emissions factor, TAM = typical animal mass, N<sub>ex</sub> = country- or region-specific N excretion rates, NA = not available. CARB (2011) estimates for liquid systems include indirect N<sub>2</sub>O emissions, which account for the non-zero value for anaerobic lagoons. The 2009 California deep pit estimate was used as the whole barn value.

### ***Emissions from Dairy Surfaces***

Greenhouse gas emissions from dairy surfaces depended on whether the surface was soil (corrals), soil covered with bedding (loafing pens), or paved (hardstandings and barn floors) and on the frequency with which accumulated manure was removed. The spatial and temporal heterogeneity in moisture content of the soil surfaces contributed to a wide range in CH<sub>4</sub> emissions, including positive and negative fluxes, and some of the highest N<sub>2</sub>O emissions of any source area (tables 6 and A5). Emissions from hardstandings and barn floors were much lower (Table 6); N<sub>2</sub>O emissions were near zero.

In corrals and loafing pens, CH<sub>4</sub> fluxes ranged over several orders of magnitude; the mean was  $0.41 \pm 68$  g hd<sup>-1</sup> d<sup>-1</sup> or  $-9.5 \pm 21$  kg ha<sup>-1</sup> d<sup>-1</sup>, where the negative sign indicates CH<sub>4</sub> uptake (Table 6). The mean values had different signs due to the difference in the studies included in the calculation (i.e., emissions measurements available in g hd<sup>-1</sup> d<sup>-1</sup> or kg ha<sup>-1</sup> d<sup>-1</sup>). Methane emissions measured from sandy and sandy loam soils using large-footprint techniques [sample mast by Leytem et al. (2011); open path Fourier transform infrared spectroscopy (FTIR) by Bjorneberg et al. (2009), and SF<sub>6</sub> tracer by Kaharabata et al. (2000)] were one to two orders of magnitude greater than emissions measured from loam to clay loam soils using dynamic chambers (Borhan et al. 2011a, 2011b). Methane can be readily oxidized in well-drained soil (Conrad 2007), thus sandy soils are generally thought to have lower CH<sub>4</sub> emissions potential than more finely textured soils. Pen conditions such as manure accumulation and high moisture, however, can slow gas diffusion and favor methanogenesis over CH<sub>4</sub> oxidation. The large- and small-footprint measurement techniques had different ranges ( $-222$ – $653$  g hd<sup>-1</sup> d<sup>-1</sup> and  $5$ – $22$  g hd<sup>-1</sup> d<sup>-1</sup>, respectively; Table A5), but the means were of the same order of magnitude ( $-6.1 \pm 124$  g hd<sup>-1</sup> d<sup>-1</sup> and  $10.2 \pm 5.8$  g hd<sup>-1</sup> d<sup>-1</sup>, respectively). As with manure management systems, small chambers deployed for shorter time periods may miss hotspots (both sources and sinks) that are better integrated into large-footprint measurements.

Corral CH<sub>4</sub> emissions were negative in 4 of 17 cases. Negative fluxes occurred in late summer when the soils were dry or in winter when the soils were cold or frozen (Table A5). Methane emissions were not correlated with either temperature or precipitation. Manure accumulation in the corrals was likely one of the most important factors driving positive CH<sub>4</sub> fluxes. Borhan et al. (2011b) measured greater CH<sub>4</sub> emissions from a dry lot corral holding heifers than from loafing pens measured at the same time ( $11$  vs.  $3.1$  g hd<sup>-1</sup> d<sup>-1</sup>, respectively; Table A5). The open lots likely had a greater influx of relatively fresh organic matter and localized high-moisture urine patches, facilitating production of anaerobic conditions and methanogenesis.

Nitrous oxide emissions from corrals were measured by three studies (Borhan et al. 2011a, 2011b; Leytem et al. 2011) and ranged over three orders of magnitude (Table 6). Corrals were the greatest source of N<sub>2</sub>O considered on a per-head basis with a mean of  $5.53 \pm 2.54$  g hd<sup>-1</sup> d<sup>-1</sup> (Table 6). The highest emissions were measured in late spring, when a combination of warmer temperatures and moist soils promoted N<sub>2</sub>O production (Table A5). The lowest values were < 50% of the highest emissions and occurred in late fall and winter (Table A5). As in the case of CH<sub>4</sub> emissions, the loafing pen measured by Borhan et al. (2011b) had lower N<sub>2</sub>O emissions rates than those of the corral ( $0.53$  vs.  $0.73$  kg ha<sup>-1</sup> d<sup>-1</sup> or  $0.60$  vs.  $2.41$  g hd<sup>-1</sup> d<sup>-1</sup>, respectively), likely for the same reasons discussed above (less manure accumulation and drier conditions in the loafing pen than in the corral). Methane and N<sub>2</sub>O fluxes from corrals were weakly positively correlated ( $r^2 = 0.28$ ,  $P < 0.05$ ) and suggest that at least a portion of the N<sub>2</sub>O fluxes were derived from denitrification, which requires the same general environmental conditions (warm temperatures, abundant labile C, anaerobic conditions) as methanogenesis.

Tier 2-predicted CH<sub>4</sub> and N<sub>2</sub>O emissions from corrals were within the standard errors of the field measurement means (Tables 7 and 8). Predicted CH<sub>4</sub> emissions were  $12.2$  g hd<sup>-1</sup> d<sup>-1</sup> (Table 7), close to the mean of  $0.41 \pm 68$  g hd<sup>-1</sup> d<sup>-1</sup> but not very informative given the large range of measurements (Table 6). Similarly, the California 2009 estimate was of the same order of magnitude as the Tier 2 estimate and

within the range of field-measured emissions rates. Even the largest field measurements were an order of magnitude smaller than the emissions from liquid manure and thus did not affect assessment of the greatest sources of CH<sub>4</sub> on California dairies.

The predicted N<sub>2</sub>O emissions rate was 4.8 g hd<sup>-1</sup> d<sup>-1</sup>, close to the field measurement mean of 5.5 ± 2.5 g hd<sup>-1</sup> d<sup>-1</sup> (Table 8). The N<sub>2</sub>O data were limited to three studies: one in Idaho (Leytem et al. 2011) and two in Texas (Borhan et al. 2011a, 2011b) (Table A5). The 2009 California estimate (Table 8) was similar to the other values, suggesting that the corrals may be the largest N<sub>2</sub>O emissions source on dairies. However, options to minimize N<sub>2</sub>O emissions from corrals are limited, and those that are feasible, such as minimizing standing water, are generally already in use.

Hardstandings and barn floors were concrete or paved with the exceptions of the freestall bedding in an open-sided barn measured by Borhan et al. (2011a, 2011b) and the brick corrals measured by Gao et al. (2011) (Table A6). Frequent scraping of the paved and concrete surfaces likely limited substrate availability and the occurrence of anaerobic conditions; thus CH<sub>4</sub> and N<sub>2</sub>O emissions were orders of magnitude lower than from other dairy sources (Table 6).

Methane emissions from most surfaces were < 8 g hd<sup>-1</sup> d<sup>-1</sup> with the exception of the brick hardstanding, which emitted 37 and 67 g hd<sup>-1</sup> d<sup>-1</sup> in the winter and summer, respectively (Gao et al. 2011) (Table A6). The higher CH<sub>4</sub> emissions from the brick hardstanding were likely due to less frequent scraping in this study (every 1–4 weeks vs. daily for most others, Table A6; Gao et al. 2011). Consistent with this finding, Adviento-Borbe et al. (2010) observed a significant, positive correlation between CH<sub>4</sub> emissions and manure depth on the barn floor. Hardstanding emissions did not have strong seasonal differences, likely due to the rapid drying of slurry in the summer, which limited anaerobic conditions, and to cold temperatures in the winter (Ellis et al. 2001). The Ellis et al. (2001) and Adviento-Borbe et al. (2010) studies indicate that solid surfaces that are scraped frequently are negligible sources of CH<sub>4</sub> and that manure accumulation is the primary factor affecting CH<sub>4</sub> emissions from hard dairy surfaces. Nitrous oxide emissions from hardstandings and barn floors were near zero; the maximum of 0.3 g hd<sup>-1</sup> d<sup>-1</sup> is from freestall bedding (Table A6; Borhan et al. 2011a). Daily scraping and dry bedding limited conditions conducive to nitrification, denitrification, or both.

IPCC default emissions factors for dry lots include paved and unpaved surfaces, such as the hardstandings and barn floors considered here. This study calculated the predicted CH<sub>4</sub> emissions rates separately for the brick hardstanding and the other surfaces to account for regional differences in dairy management. Predicted CH<sub>4</sub> emissions were 8.1 g hd<sup>-1</sup> d<sup>-1</sup> for the hardstandings and barn floors, the same order of magnitude as the field mean of 3 ± 3 g hd<sup>-1</sup> d<sup>-1</sup> (Table 6). Predicted CH<sub>4</sub> emissions from the brick hardstanding were even lower at 2.4 g/hd/y, less than 5% of the average from Gao et al. (2011) (Table 6). This result is likely due to use of the default Asian *B<sub>o</sub>* and *V<sub>S</sub>* values, which may not represent the modern Holstein dairy that was measured (Gao et al. 2011). Thus, the IPCC Tier 2 approach is adequate for calculating CH<sub>4</sub> emissions from most dairy surfaces, but it may need to be refined for the rapidly modernizing Asian dairy industry.

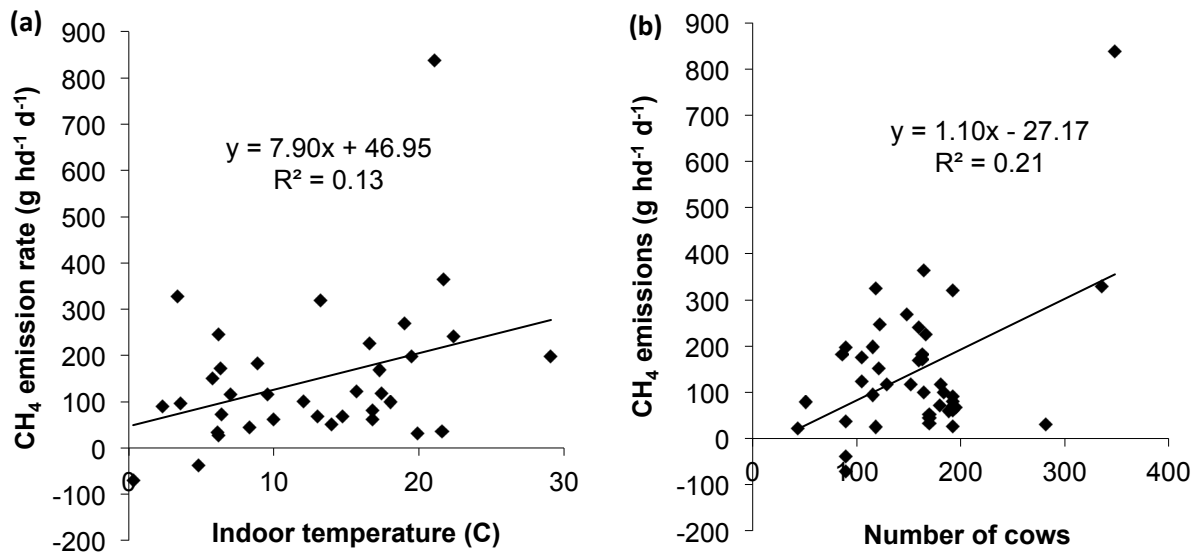
Predicted N<sub>2</sub>O emissions for the hardstandings and barn floor were the same as for the dry lot at 8.3 g hd<sup>-1</sup> d<sup>-1</sup>, three orders of magnitude larger than the field mean of 0.0012 ± 0.0002 g hd<sup>-1</sup> d<sup>-1</sup> (Table 7). Hardstandings and barn floors are not the primary manure management area for most dairies, and therefore these emissions are often excluded from local and regional estimates. More measurements of N<sub>2</sub>O emissions for hardstandings and barn floors are needed to determine if manure accumulation might increase N<sub>2</sub>O emissions to rates comparable with those predicted by the Tier 2 approach.

In summary, GHG emissions from dairy surfaces varied widely depending on the surface composition and the amount of manure accumulation on that surface. Dry lot corrals can be significant sources of N<sub>2</sub>O. Low moisture content, low temperature, and frequent manure removal appeared to limit GHG emissions from dairy surfaces.

### Emissions from Whole Barns

Six studies measured CH<sub>4</sub> emissions from whole barns and two measured N<sub>2</sub>O (Table A7). Kinsman et al. (1995) was the only study that measured an empty barn, i.e., excluded enteric emissions. All the measured barns were enclosed tiestall or freestall barns, unlike most barns in California which are typically open-sided. Mean CH<sub>4</sub> emissions (corrected for enteric emissions) were  $158 \pm 80 \text{ kg ha}^{-1} \text{ d}^{-1}$  or  $153 \pm 75 \text{ g hd}^{-1} \text{ d}^{-1}$ , intermediate between emissions from manure management and from surfaces (Table 5). Samer et al. (2012) measured the highest barn CH<sub>4</sub> emissions in summer from a barn that had the highest ventilation rate, highest number of cows, and largest cows (Table A7). The lowest CH<sub>4</sub> emissions were measured in winter from a barn with the fewest cows and a relatively high ventilation rate (Table A7) (Marik and Levin 1996). Methane emissions were usually higher in summer than in winter and were weakly positively correlated with indoor temperature (Figure 3a). Methane emissions also increased with number of cows (Figure 3b), which could be due to the greater manure accumulation associated with greater numbers of cows, or it could indicate that this study's enteric emissions correction may be an underestimate for European dairies that make up the bulk of the studies considered here. Zhang et al. (2005) observed higher CH<sub>4</sub> emissions from barns with slatted floors than from ones with solid floors. This finding may reflect the manure handling systems associated with the different floor types, because almost all the slatted floor barns had flushing or circulation systems, whereas solid floors were mechanically scraped.

**Figure 3. Whole Barn CH<sub>4</sub> Emissions (Corrected for Enteric Emissions Assuming 100 kg CH<sub>4</sub> hd<sup>-1</sup> y<sup>-1</sup>) Increase with (a) Indoor Barn Temperature and (b) Number of Cows**



Measurements of nitrous oxide emissions from whole barns in two studies yielded very different results. The mean of measurements by Samer et al. (2012) was  $60.36 \pm 1.08 \text{ g hd}^{-1} \text{ d}^{-1}$  and was the largest calculated for any dairy source. In contrast, and despite using a similar measurement technique, the mean from Zhang et al. (2005) was about 25 times less ( $2.37 \pm 0.47 \text{ g hd}^{-1} \text{ d}^{-1}$ , Table A7) and was comparable to N<sub>2</sub>O emissions from other dairy sources (Table 6). Nitrous oxide fluxes typically vary widely in space

and time as a result of the sensitivity of nitrification and denitrification to local conditions (Firestone and Davidson 1989). In this case, differences in barn conditions in the Samer et al. (2012) and Zhang et al. (2005) studies appeared to create large differences in emissions, but identifying the key factors driving these differences is difficult. Zhang et al. (2005) found that N<sub>2</sub>O emissions were generally greater in winter than summer, and no specific combinations of flooring or manure handling appeared to consistently produce higher or lower values.

The IPCC has no specific emissions factors for barns, but deep bedding and pit storage in barns are included in manure management calculations. Though few of the barns considered here had deep bedding or pit storage, they collect some manure and so this study calculated emissions under the assumption that pit storage was present (Tables 6 and 7). Predicted CH<sub>4</sub> emissions were 24 g hd<sup>-1</sup> d<sup>-1</sup>, much lower than the field mean. This discrepancy could result from an underestimation of enteric emissions in this study or an underestimation in the emissions factor of the amount of manure typically present in barns. After recalculation of the field mean CH<sub>4</sub> emissions using the IPCC (2006) default enteric emissions rate for Western Europe (320 g hd<sup>-1</sup> d<sup>-1</sup>), the field mean decreased to 124 ± 80 g hd<sup>-1</sup> d<sup>-1</sup>, still well above the predicted value. Thus, the data suggest CH<sub>4</sub> emissions from manure in barns may be as great as or greater than emissions from subfloor manure storage tanks or deep bedding, and emissions factors accounting for these sources should be developed and included in future emissions calculations.

The predicted N<sub>2</sub>O emissions rate of 0.91 g hd<sup>-1</sup> d<sup>-1</sup> was two orders of magnitude less than the field mean and less than 40% of the mean of the measurements from Zhang et al. (2005). Enteric N<sub>2</sub>O emissions are not considered by the IPCC (2006), but they are generally low (≤0.5 g hd<sup>-1</sup> d<sup>-1</sup>; Mitloehner, Sun, and Karlik 2009) and are insufficient to account for the difference between measured and predicted values. This result suggests that the current emissions factors may underestimate N<sub>2</sub>O emissions from some barns, but this possibility is highly uncertain, reflecting data from only two studies.

In California, deep pit storage is relatively rare, accounting for manure treatment of less than 5,000 hd (Table 3), so the accuracy of this estimate is not as critical as it is for other liquid systems. Additionally, most California barns are open-sided and are scraped or flushed. Thus, barns in California are likely small sources of GHG emissions, provided they do not have deep pit storage and manure is removed frequently (as discussed for paved dairy surfaces).

### ***Emissions from Whole Dairies***

Comparatively fast and non-intrusive field measurements can be made at whole dairies. These measurements and their accuracy are of interest for monitoring emissions of regulation compliance. Two studies have measured CH<sub>4</sub> emissions from a total of 18 whole dairies in the Netherlands and Canada (Tables 4 and A8). Measurement techniques included a mobile tunable diode laser sensor for a single drive-by measurement of each dairy (Hensen et al. 2005 and 2006) and an open path laser deployed for a few weeks at each dairy (McGinn and Beauchemin 2012). No studies measured N<sub>2</sub>O from whole dairies.

When corrected for enteric emissions, mean CH<sub>4</sub> emissions from all other sources on the dairies were 309 ± 96 g hd<sup>-1</sup> d<sup>-1</sup>, with a range of -209–1005 g hd<sup>-1</sup> d<sup>-1</sup> (Table 6). Winter emissions tended to be lower than summer emissions. The IPCC has no explicit Tier 2 approach for calculating emissions from entire dairies, but the equivalent is summing the predicted emissions from sources found on the dairies included in these studies. This study used default manure characteristic values for western Europe ( $VS = 1861.5$  g hd<sup>-1</sup> y<sup>-1</sup>,  $B_o = 0.24$  m<sup>3</sup> kg<sup>-1</sup>) and cool MAT *MCFs* for corrals (1%), manure piles (0.5 %, value for static pile), slurry tanks (10 %), and barns (3 %, for deep bedding or pit storage) (IPCC 2006) to calculate a whole dairy emissions factor of 117 g hd<sup>-1</sup> d<sup>-1</sup>. This factor is about 40 % of the field measurement mean (Table 6). The more general Tier 1 estimate of the CH<sub>4</sub> emissions rate from manure management, which

includes all sources except enteric emissions, is  $\sim 57 \text{ g hd}^{-1} \text{ d}^{-1}$  for western Europe, or less than 20% of the measured value. The field measurement mean may be higher than the calculated annual average because the measurements were made mostly in early summer (Table A8) when emissions would be expected to be relatively high. Recognizing this mismatch, Hensen et al. (2005, 2006) calculated month-specific, rather than annual,  $\text{CH}_4$  emissions rates using country-specific emissions factors and the Farm Greenhouse Gas model. Their calculated emissions rates were generally within 20–50 % of the measured emissions and were about evenly split between being too high and too low. Because the month-specific calculations addressed the seasonal bias that the Tier 2 approach could not, they were slightly more accurate.

For comparison, this study calculated a whole dairy value for California by summing the emissions from anaerobic lagoons, corrals, and solid manure piles, yielding a total value of  $816 \text{ g hd}^{-1} \text{ d}^{-1}$  (Table 7). This value is comparable to that of the four dairies with the highest measured emissions (Table A8); 95% of these emissions came from anaerobic lagoons. This finding again highlights the large GHG emissions associated with anaerobic lagoons.

### ***Implications for California Dairies***

None of the studies included in the analysis above were done in California, and variables critical to development of representative emission factors, such as  $B_o$  and  $VS$  in equation 1, have not been determined for California dairies. Assuming model values accurately represent California dairies, GHG emissions from liquid manure management systems, particularly  $\text{CH}_4$  from anaerobic lagoons, which were the highest total and per-cow source, represent the greatest opportunity for GHG emissions reduction from California dairies.

Before mitigation options and potential are discussed, several limitations of this comparison must be acknowledged. One weakness of the above analysis is the lack of emissions model parameters specific to California dairies. These parameters would reduce the error and uncertainty in estimates. Another weakness is the lack of field measurements to confirm emissions estimates. Model parameterization and estimate confirmation both require field measurements. Lack of field measurements in California is one of the greatest sources of uncertainty in the GHG inventory.

Gerber et al. (2013) report several options for  $\text{CH}_4$  reduction from anaerobic lagoons and slurry systems:

- emplacement of a natural or induced crust,
- sealed containment with flaring of gas (simple anaerobic digestion),
- conversion to an advanced anaerobic digestion system with biogas collection for energy production, and
- reduction of the volume of manure in liquid storage systems while maintaining milk production.

The goal of the first three options is to convert  $\text{CH}_4$  to climate-neutral  $\text{CO}_2$ .

#### **Crusts**

Crusts may provide an oxidative environment in which methanotrophs consume  $\text{CH}_4$  as it diffuses out of the liquid manure. These same conditions, however, may promote  $\text{N}_2\text{O}$  production. No studies have been made of the effect of crusts on GHG emissions from California liquid manure systems. Therefore, the efficacy of this approach is unknown and likely depends on climate and management practices.

## Anaerobic Digestion

CARB (2011) GHG estimates indicate that the total CH<sub>4</sub> and N<sub>2</sub>O emissions per head from anaerobic digestion is about 10% of the emissions from anaerobic lagoons (Table 3). Other benefits of anaerobic digesters include:

- production of energy for on-farm use (and possible purchase by utility companies) through the combustion of biogas (largely CH<sub>4</sub>),
- decreased manure volume (which decreases subsequent handling costs and energy),
- improved fertilizer value of the effluent (discussed below),
- elimination of some pathogens (Kearney et al. 1993) and weed seeds (Johansen et al. 2013) from the manure, and
- decreased odors (USEPA 2011a; Massé et al. 2011; Camarillo et al. 2012).

The USEPA (2011b) estimated that adoption of anaerobic digesters by all the California dairies for which this technology is feasible (those with liquid manure management systems and > 500 hd, or approximately 900 farms with 1.35 million hd) could reduce California's CH<sub>4</sub> emissions by 7.7 Tg CO<sub>2</sub>e y<sup>-1</sup>, or 92% of the total CH<sub>4</sub> emissions from dairy manure management. In addition, these anaerobic digesters would be capable of producing more than 2 million MWh y<sup>-1</sup> (USEPA 2011b). According to the GHG equivalency calculator at [www.epa.gov](http://www.epa.gov), this energy production offsets an additional 1.41 Tg CO<sub>2</sub>e (considering CO<sub>2</sub> emissions from energy production).

Several obstacles have been encountered with implementation of anaerobic digesters, which has led to their abandonment and to reluctance of dairies to invest in them (CEC 2009; Camarillo et al. 2012). These obstacles include:

- issues with connectivity to utilities and reimbursement for excess energy production,
- long and sometimes complicated permitting processes,
- high construction and maintenance costs,
- issues related to design (leaking, corrosion, clogging, complexity),
- operation below maximum efficiency, and
- production of ozone precursors (particularly NO<sub>x</sub> and VOCs) during the combustion of biogas collected from the digester.

Although NO<sub>x</sub> and VOC emissions are not a problem in all areas, they are regulated in many California counties and are costly to mitigate. Current work is focused on improving the economic viability of anaerobic digesters through legislation to streamline the permit process and facilitate cooperation of utilities to increase economic benefits. Other work is focused on improving anaerobic digester design for better performance and decreased emission of undesirable gases (Ward et al. 2008).

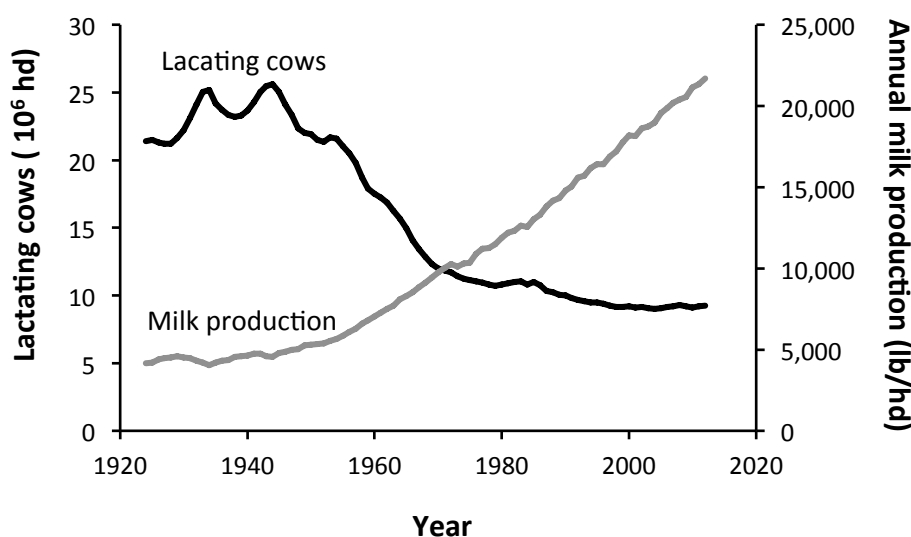
## Decreased Volume of Manure in Liquid Systems

Decreasing the amount of manure in liquid management systems could be achieved by managing more manure in solid form or decreasing the total number of cows while retaining total milk production. Managing more manure in solid form would be feasible if dairies were still being constructed, but recent economic stresses have instead decreased the number of dairies (CDFA 2013). Thus, doing so would involve a major renovation of dairy infrastructure. A full life-cycle assessment (LCA, discussed below) would also be required to ensure that managing manure primarily as a solid would result in a net decrease in GHG emissions.



Decreasing the number of cows decreases both enteric emissions and manure production, but not necessarily at the expense of total milk production. Increasing milk production per cow increases the milk production per amount of GHG emissions from manure management and enteric fermentation. The number of cows in the United States has steadily declined as milk production has increased (Figure 7, <http://www.nass.usda.gov/>). Between 1944 and 2007, U.S. milk production efficiency increased such that, per billion kg of milk produced, the number of lactating cows decreased from 414,800 to 93,600, N excretion decreased from 17.47 Gg N to 7.91 Gg N, and CH<sub>4</sub> emissions decreased from 61.8 Gg CH<sub>4</sub> to 26.8 Gg CH<sub>4</sub> (Capper et al. 2009). Variation in the amount of milk produced per lactation and per cow over her lifetime remains large, indicating the high potential for genetic selection to further increase milk production efficiency while decreasing management costs and GHG emissions (Knapp et al. 2011).

**Figure 4. Decreasing Number of Milk Cows and Increasing Production per Head in the United States between 1924 and 2013**



Source: <http://www.nass.usda.gov/>.

#### **Life-cycle Assessments of Dairy Manure Management**

Quantifying the sinks and sources of C and N related to manure management on California dairies is critical for identifying management practices that reduce GHG emissions as well as for parameterizing LCAs. LCAs are powerful tools for guiding GHG reduction efforts in that they calculate the GHG emissions associated with all steps of production of a product, typically up to the point that the product leaves the farm (“cradle-to-gate”). They are especially important for determining the net effect of GHG management strategies, ensuring that they do not actually result in increased GHG emissions or emissions of other air pollutants. Parameterizing emissions models for California dairies is under way (Salas et al. 2008; CARB 2011), but constraining emissions factors and their relationship to manure management approaches will enable more confident, accurate estimates of emissions from dairies.

Numerous models have been developed for dairy and agricultural systems on the basis of varying combinations of emission factors, empirical relationships, and process-based calculations (see Table 1 in Li et al. 2012 for a summary of those developed for manure management). The utility of models is dependent on the availability of high-quality data for parameterization. Such data are lacking in most areas of California.

Manure-DNDC (Li et al. 2012) is a modification of the biogeochemical, process-based Denitrification-Decomposition (DNDC) model and links the core of DNDC to a virtual animal farm. Li et al. (2012) developed Manure-DNDC to predict GHG and NH<sub>3</sub> emissions from manure in storage and field application. The Canadian government has commissioned an integration of animal (dairy) emissions models and Manure-DNDC, which E. Kebreab and W. Salas (Applied GeoSolutions) are conducting. An integrated swine model for estimating air emissions from both animals and manure management is also in its final stages (E. Kebreab). Further work is necessary for integrating emissions models for dairy cow enteric emissions and manure management under conditions in California using local data wherever possible.

The DairyGHG model (Rotz et al. 2010) can perform partial LCA (cradle to farm gate) of dairy farms. It was used to compare the GHG costs of milk production for a hypothetical Pennsylvania dairy (Chianese et al. 2009a,b,c) as well as two hypothetical California dairies (Rotz et al. 2010). DairyGHG simulated significant differences in global warming potential for dairies utilizing different management practices. However, the lack of field data for comparison with, or calibration of, the California dairy simulations is problematic and indicates a significant knowledge gap.

The Integrated Farm System Model (ISFM; Rotz et al. 2011) was developed to track environmental impacts (P loss, NH<sub>3</sub> and hydrogen sulfide emissions) and economic costs for typical farming systems. It has since been expanded to account for GHG emissions. As with other models, calibration of the ISFM for California systems is limited, but this model is one of the few options for concurrently analyzing the economic and environmental impacts of management practices designed to decrease GHG emissions.

The selection of GHG mitigation practices requires consideration of the net effects of the practice on GHG and other criteria air pollutants. This analysis is best achieved through life-cycle assessments of various types of emissions, other environmental impacts, and economic impacts. Field and laboratory data necessary for constraining and confirming these assessments are nearly absent for California. There is no question that decreasing emissions from anaerobic lagoons will have a significant impact on total GHG emissions, but the absolute decrease in emissions is uncertain without well-constrained models to determine it.

## **LAND APPLICATION OF MANURE**

Dairy manure that does not volatilize, decompose, or evaporate during on-dairy management is eventually applied to agricultural land as a fertilizer, and liquid manure can be used for irrigation. Agricultural soils typically emit little CH<sub>4</sub> with the exception of rice cultivation. This report focuses on the sequestration of C in the soil, which decreases the emission of CO<sub>2</sub> and CH<sub>4</sub> associated with manure decomposition. Nitrous oxide emissions from land application of manure are discussed by Culman et al. (2014).

Using manure as fertilizer rather than inorganic fertilizers has benefits and challenges. Benefits include decreasing or eliminating entirely the need for inorganic fertilizer and increasing soil C. Compared with manure, synthetic fertilizers have a higher energy cost and global warming potential associated with their production and application, assuming similar transport distances (Wiens et al. 2008; Eagle and Olander 2012).

Long-term intensive cultivation and inorganic fertilization of agricultural fields has significantly reduced soil organic C content by more than 1000 Tg in the United States (Kern and Johnson 1993). This soil C was largely released to the atmosphere as CO<sub>2</sub>. Because soil C was not replenished, agricultural soils are a net source of greenhouse gases. As described below, manure additions can increase soil C storage through

direct (manure) and indirect (plant growth) inputs (Eagle and Olander 2012). Increased soil organic matter decreases bulk density, increases water and nutrient holding capacity, and improves aggregation, all of which can promote plant growth (Brady and Weil 2010). Most important to this report, soil organic matter also sequesters C from the atmosphere.

One of the greatest challenges of fertilization with manure is that a large portion of the N in manure is in the organic form and is released more slowly and less predictably to the soil through decomposition than through inorganic fertilization (Van Kessel and Reeves 2002). Because of this slow release, farmers are less able to match N availability with crop demand, which is typically non-linear through the growing season (Gastal and Lemaire 2002). Manure is also a composite fertilizer containing multiple nutrients that are added together. Thus, the optimum application rate for N may under- or overapply P and K, leading to nutrient deficiencies or pollution (Van Kessel and Reeves 2002). Lastly, due to its solids content, liquid manure may require application techniques different from those for inorganic fertilizer dissolved in irrigation water (Pfoest et al. 2001).

### ***Carbon Storage in Agricultural Fields through Manure Addition***

Manure addition has been shown to increase soil C storage in California (Poudel et al. 2001; Kong et al. 2005; Kong et al. 2009) and in fields in other regions with a Mediterranean climate (Aguilera et al. 2013a). In 10 different cropping systems after 10 years of management at the Long-Term Research of Agricultural Systems (LTRAS) site, organic management (legume cover cropping and composted manure amendment) resulted in an average SOC sequestration rate of  $0.56 \text{ Mg C ha}^{-1} \text{ y}^{-1}$  in the top 15 cm of the soil, whereas most other management systems were small sources or significantly smaller sinks (less than  $\pm 0.06 \text{ Mg C ha}^{-1} \text{ y}^{-1}$ ) (Kong et al. 2005). Much of the soil C gain in the organic fields was due to high C inputs ( $89.6 \text{ Mg C ha}^{-1}$  over the 10 years). Soil C accumulation increased the number and stability of macroaggregates in soils, which may lead to longer C residence times in soils. A comparison between fields revealed the importance of amendment quality in soil C storage capacity. It also revealed that fields with manure amendment were more efficient at storing C than similarly managed fields without manure amendment (Kong et al. 2005).

A similar experiment at the Sustainable Agricultural Farming Systems (SAFS) research site near Davis, California, showed that fields organically farmed for 10 years using poultry manure increased soil C concentrations in the top 15 cm of soil compared to conventional fields ( $12.6 \text{ g kg}^{-1}$  vs.  $9.6 \text{ g kg}^{-1}$ , respectively) (Poudel et al. 2001). Assuming a soil bulk density of  $1.2 \text{ g cm}^{-3}$  (the average for the nearby LTRAS site; Burger et al. 2005), the difference in C content between synthetic fertilization and manure amendment was  $5.4 \text{ Mg C ha}^{-1}$ , equivalent to  $0.54 \text{ Mg C ha}^{-1} \text{ y}^{-1}$  or  $2 \text{ Mg CO}_2\text{e ha}^{-1} \text{ y}^{-1}$ . Soil C can increase rapidly; at the Russell Ranch experimental site, also near Davis, organic management (composted dairy manure amendment) increased SOC by approximately  $5 \text{ Mg C ha}^{-1}$  compared with conventional and low-input management one year after conversion, regardless of tillage (Kong et al. 2009).

The sequestration rates reviewed above are within the range found in other Mediterranean agricultural systems, where C sequestration rates ranged from  $0.5$  to  $2.3 \text{ Mg C ha}^{-1} \text{ y}^{-1}$  following manure addition (Aguilera et al. 2013a). Carbon sequestration in fields with livestock slurry amendment, however, showed no significant difference from conventional fields. Though this finding was based on only three studies, it is consistent with results from other regions. Slurry typically has relatively high mineralized N content and low C content which promotes microbial growth and the decomposition of soil organic matter and release, rather than storage, of  $\text{N}_2\text{O}$  (Aguilera et al. 2013a).

The California-based field studies described above provide a useful starting point for understanding C sequestration in agricultural fields amended with manure, but they represent a small cross-section of

typical agricultural practices in the state for a limited geographic area on only three different soil series. Carefully controlled and replicated field studies in a wider range of the state's bioclimatic zones and on more soil types are needed to expand this analysis.

**Impacts of Manure Type**

Manure composition and form can have significant effects on N<sub>2</sub>O production and C storage, particularly in Mediterranean climates where soil processes are sensitive to changes in soil moisture associated with slurry application (Aguilera et al. 2013a,b). Anaerobic digester output has received less attention than solid and liquid manures, but its availability in California could increase. As of 2009, at least 10 California dairies had anaerobic digesters as a result of the Dairy Power Production Program undertaken by the California Energy Commission (CEC) as part of Senate Bill 5X (CEC 2009). In September 2012, California passed California Senate Bill 1122 (SB1122), which directs the California Public Utilities Commission to require the utility industry to produce at least 250 megawatts from bioenergy operations that commence on or after June 1, 2013. Much of this mandate is expected to be met with biogas production from anaerobic digesters utilizing municipal waste, food processing waste, and livestock manure.

Few studies have compared the GHG emissions from land application of anaerobic digestate and other manure types. Total GHG emissions (CH<sub>4</sub> and N<sub>2</sub>O) from pilot-scale manure storage and land application in Austria were lower for anaerobically digested slurry than for untreated slurry, separated slurry, slurry covered with straw while in a tank, and aerated slurry (Amon et al. 2006). Similar findings were found on a xeric, irrigated agricultural field in eastern Washington (Collins et al. 2011), where cumulative N<sub>2</sub>O emissions were greater for dairy slurry- and digested effluent-amended fields than for fields treated with synthetic fertilizer. The total global warming potential, however, was lowest for those amended with the digested effluent (Table 10) and was approximately 20 kg CO<sub>2</sub>e ha<sup>-1</sup> lower than for fields amended with synthetic fertilizer, and 25–130 kg CO<sub>2</sub>e ha<sup>-1</sup> lower than for fields amended with untreated liquid manure.

**Table 10. Net Global Warming Potential (kg CO<sub>2</sub>e ha<sup>-1</sup>, Mean ± Standard Error, Considering N<sub>2</sub>O and CH<sub>4</sub>) of Corn Fields near Prosser, Washington, with Different Amendments Over Two Study Seasons**

Treatment	2007	2008
Fallow	63.3 ± 17	38.0 ± 10
Unfertilized	51.4 ± 22	86.9 ± 15
Synthetic fertilizer	114.0 ± 12	225.1 ± 40
Digested fiber	115.0 ± 35	62.5 ± 17
Digested effluent	90.6 ± 22	192.5 ± 69
Liquid manure	219 ± 18	217.7 ± 22

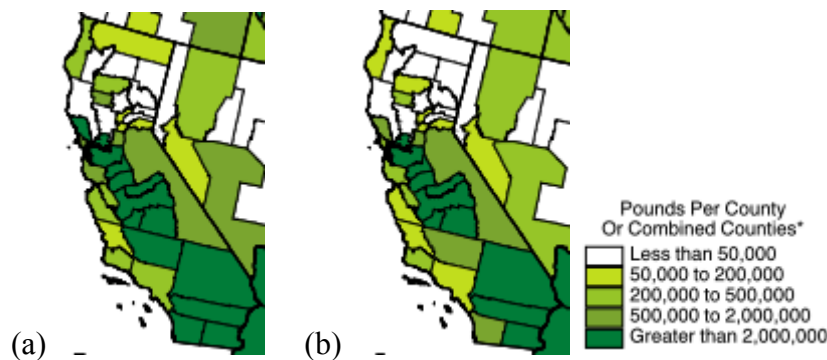
Source: Table 3 in Collins et al. (2011).

The consistency of results in the Amon et al. (2006) and Collins et al. (2011) studies suggests that land application of the outputs from anaerobic digestion may lower GHG emissions and maintain or improve crop yields under conditions and practices common in California. Very little field work has been done in California to quantify GHG emissions related to land application of manure and the products of anaerobic digestion. Using rates estimated outside California, application of anaerobic digestate, rather than synthetic fertilizer or liquid manure, could decrease GHG emissions by 20-130 kg CO<sub>2</sub>e ha<sup>-1</sup> y<sup>-1</sup>.

### Limitations to Land Application of Manure

Over the last few decades, the increasing consolidation of the California dairy industry has resulted in more large CAFOs and their concentration in the San Joaquin Valley and other areas (Kellogg et al. 2000; CDFA 2013). This concentration may create a water pollution issue as more manure is produced in an area with insufficient land for application at levels that do not compromise water quality. The amount of N and P in livestock manure by farm and by county in the United States has been estimated and compared with the estimated land available (permanent pasture and agricultural land associated with 24 crops) to receive that manure (Kellogg et al. 2000). The southern valleys of California stand out as regions of high excess N and P, assuming no manure export from farms (maps 28 and 30, Kellogg et al. 2000, reproduced in Figure 5a,b).

**Figure 5. Excess Manure N (a) and P (b) by County in 1997 in California, Assuming No Manure Export from Farms**

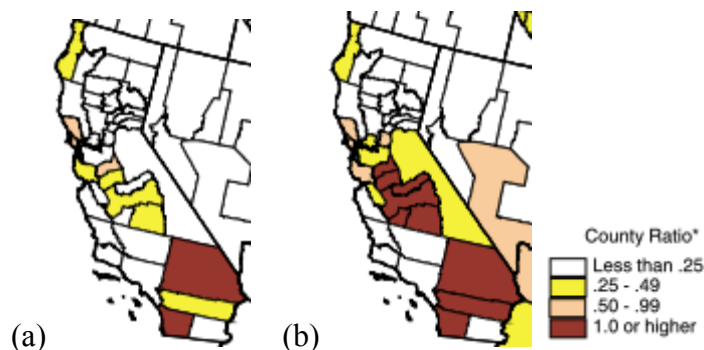


Source: Maps 28 and 30, respectively, in Kellogg et al. (2000).

Note: Includes all livestock.

Considered by county and assuming a more realistic scenario in which livestock operations could dispose of excess manure on other land within the county, the ratios of nutrient production to nutrient assimilation capacity for both N and P were less than 1 (where > 1 indicates excess nutrients and < 1 indicates crop assimilation of all the manure nutrients) for almost all California counties in 1997. San Bernardino and San Diego counties had the potential for excess N (Map 35 in Kellogg et al. 2000, reproduced in Figure 6a); several more counties had the potential for excess P (Map 37 in Kellogg et al. 2000, reproduced in Figure 6b). Total farm-level excess N was estimated as ~61,000 Mg in California in 1997 (Kellogg et al. 2000). Excess N addition to soils increases N<sub>2</sub>O emissions.

**Figure 6. Ratio of Manure N Produced to Assimilation Capacity (a) and Ratio of Manure P Produced to Assimilation Capacity (b) in 1997 in California**



Source: Maps 35 and 36, respectively, in Kellogg et al. (2000).

The above analysis may not accurately reflect the current extent of nutrient excess in California due to the more recent implementation of the NRCS CNMP program, which limits manure N application to 140% of crop N needs. This reduction is significant and necessitates more off-farm options for manure disposal (Owen and Silver, in prep). Additionally, other constituents of some manure (e.g., hormones, antibiotics, pathogens, salts) have the potential to affect soil and water systems.

#### ***Application of Manure to Rangeland***

Though most livestock manure is eventually applied to agricultural fields, some is applied to rangelands. This practice may become more common as both a means of manure disposal and a global warming mitigation strategy. Manure application rates to agricultural fields on dairies are now limited by estimates of plant nutrient demand. Thus, application to rangelands may be a viable alternative and may offer the co-benefits of increasing forage production and soil carbon sequestration (see DeLonge, Owen, and Silver 2014).

#### ***Modeling Land Application of Manure***

Biogeochemical models can be used to estimate the net effect of different management practices on GHG emissions and soil C and N, provided they are adequately parameterized for the soil and crop types of interest. This study focuses on models of soil C. De Gryze et al. (2010) calibrated the DAYCENT model with data from four long-term agricultural experiments. According to their estimates, soil C sequestration is much greater in organic fields than in conventional fields, regardless of tillage (Table 11). These estimates suggest that switching from conventional management to manure application could decrease total GHG emissions by  $> 4 \text{ Mg CO}_2\text{e ha}^{-1} \text{ y}^{-1}$  when soil C is considered. Estimates from DAYCENT generally agreed with field measurements, with variability introduced by inter-annual climate variability and by factors related to crop growth that are not included in the model. They found that manure amendments had lower potential GHG emissions than inorganic fertilizers.

**Table 11. DAYCENT-Derived Estimates for Change in Soil Organic C in the Top 1.5 m of Soil ( $\Delta$ SOC) and Net Soil GHG Flux for the Long-Term Research on Agricultural Systems Study Site**

Management	$\Delta$ SOC (kg C ha <sup>-1</sup> y <sup>-1</sup> )	Net greenhouse gas flux (kg CO <sub>2</sub> e ha <sup>-1</sup> y <sup>-1</sup> )
<i>Standard tillage</i>		
Synthetic	95 ± 46	1081 ± 192
Low input	315 ± 46	9 ± 192
Organic	1324 ± 46	-3496 ± 192
<i>Conservation tillage</i>		
Synthetic	47 ± 87	1182 ± 391
Low input	321 ± 87	-192 ± 391
Organic	1279 ± 87	-3349 ± 391

Source: Table 5 in De Gryze et al. (2010).

Note: Positive values of  $\Delta$ SOC indicate an increase of C in the soil, whereas positive values for net soil GHG flux indicate loss to the atmosphere. Synthetic = synthetic fertilizer; low input = less synthetic fertilizer and cover cropping; organic = poultry manure fertilizer and cover cropping.

Alternative management practices were modeled for eight crops grown in ten counties in the Central Valley of California to determine their effect on crop yields and GHG emissions (De Gryze et al. 2011). Using manure instead of synthetic fertilizer decreased yields by less than 4%, in most cases, while decreasing GHG emissions by 1.16 Mg CO<sub>2</sub>e ha<sup>-1</sup> y<sup>-1</sup> in the Sacramento Valley and 0.5 Mg CO<sub>2</sub>e ha<sup>-1</sup> y<sup>-1</sup> in the San Joaquin Valley. A large portion of the decrease in GHG emissions with the switch from synthetic fertilizer to manure was due to accumulation of soil organic C, which could be reversed with increased tillage. These estimated changes in emissions did not include changes associated with decreased synthetic fertilizer production or changes in fuel use with adoption of the alternative management practices. Detailed manure characteristics used in the model were not specified; results from LTRAS were used to constrain the model (De Gryze et al. 2010), but this study used composted poultry manure, which may not be applicable to dairy manure-amended fields.

In summary, modeling tools that can estimate GHG emissions from land application of manure have been developed and are being refined and parameterized for California. Modeling results suggest that using manure rather than synthetic fertilizer could result in a decrease of 0.5-4 Mg CO<sub>2</sub>e ha<sup>-1</sup> y<sup>-1</sup>. Complete life-cycle assessment is needed to determine the total GHG reduction associated with this switch. Such assessment would address fuel use for land management and manure transportation, synthetic fertilizer production, and the emissions of other pollutants of concern such as NH<sub>3</sub>. Model application to California is limited by the lack of field data to constrain the emissions factors and process models at their core. California's diverse agriculture and soils, Mediterranean climate, and extensive use of anaerobic lagoons and flood irrigation are substantial challenges that the models must overcome.

### **Implications for California Agriculture**

The results discussed above suggest that GHG emissions from land application of solid manure or liquid effluent from anaerobic digesters are lower than from fields treated with synthetic fertilizers. Modeling results suggest that using manure rather than synthetic fertilizer could result in a decrease of 0.5-4 Mg CO<sub>2</sub>e ha<sup>-1</sup> y<sup>-1</sup>. Unfortunately, none of the studies measured or modeled GHG emissions from the most common land application practice on California dairies: furrow irrigation with diluted lagoon effluent. This practice may increase emissions of both CH<sub>4</sub> and N<sub>2</sub>O.

Field studies of California's typical dairy agricultural fields are needed to improve model estimates and identify practices that decrease GHG emissions. Particularly critical are long-term agricultural experiments that are representative of practices on dairy-associated fields, including furrow irrigation with diluted liquid manure from anaerobic lagoons.

## KEY FINDINGS AND RESEARCH PRIORITIES

The literature review described above identified significant opportunities for GHG mitigation:

- **Livestock manure, particularly dairy manure stored in anaerobic lagoons, is likely to be the largest source of GHG emissions from California agriculture.** Manure emissions potentially could be decreased more than 90% if lagoons are capped (and their methane emissions flared), farms transition to anaerobic digesters, or both.
- **Decreasing the amount of manure treated in liquid systems, whether by changing manure management practices or decreasing the number of cows, would also decrease total GHG emissions.** Anaerobic lagoons, the most common on-farm dairy manure storage practice in California, have more than 10 times the global warming potential of solid manure piles. Milk production has been steadily increasing despite a national decrease in cow population. Future advances in breeding and feed management could further reduce the amount of manure produced for the same or increased milk production.
- **Land amended with manure had lower emissions than fields managed with synthetic fertilizer and had consistently greater soil C content. Compared with other manure management practices, land application decreased GHG emissions associated with manure decomposition.** Research from outside California suggests that anaerobic lagoon effluent may result in greater emissions following field application compared with solid manure and anaerobic digester effluent, yet both types of effluent have lower emissions than fields managed with synthetic fertilizer.

We identified several key research needs:

- **California-specific measurements of GHG fluxes from manure management are needed, particularly from anaerobic lagoons on dairies, anaerobic digesters, fields with furrow irrigation and fertilization, and land-applied manure and anaerobic digester effluent.** No GHG emissions measurements from manure management or land application of dairy manure in California have been published. These data are needed to improve GHG emissions estimates and to identify best management practices. Long-term measurements of soil C and N content, CH<sub>4</sub> and N<sub>2</sub>O emissions, and crop yields (including pasture productivity) in conjunction with local climate data, management data, and soil characteristics are especially critical.
- **Comparisons of field measurement techniques are needed to facilitate comparisons across studies and scales of analysis.** This literature review revealed wide variations in GHG emissions measurements from some sources—variations that may be the result of technique, rather than differences in actual emissions. Scientists must critically evaluate methodologies to identify the measurement techniques that most accurately quantify GHG emissions from heterogeneous manure storage systems and from areas where manure has been applied.
- **Process-based models are increasingly important in informing policy decisions, but must be parameterized for the range of manure management practices, crops, and irrigation systems found in California.** The lack of field and laboratory research on GHG sources and sinks, and the mechanisms controlling these dynamics in California severely limit the utility of models. Integration of animal, soil, and, ideally, crop biogeochemical models is needed at local and regional scales.
- **Surveys of manure management practices are needed for all major dairying regions of California.** These surveys would expand the work of Meyer et al. (2011). Accurate emissions estimates require better estimates of (1) the amount of manure C and N under the various



management practices, (2) the spatial distribution of these practices across California, and (3) the length of storage and ultimate fate of manure. These inputs are critical for identifying trends in manure management in the dairy industry as well as for quantifying California's excess manure nutrient production, which may affect land application practices and have economic impacts on farmers.

The California dairy industry is already at the forefront of environmental sustainability innovations, and its successful mitigation of GHG emissions, coupled with economic growth, can serve as an example to other dairy industries across the world. More emissions data and improved models, however, are needed to accurately quantify this achievement and optimize energy, crop, and economic returns from manure management.

## APPENDIX A: DATA FOR DAIRY EMISSIONS STUDIES

Table A1. Study location, measurement technique, and climate data for the 32 studies of dairy emissions.

Study	Measurement technique <sup>a</sup>	Gas analysis technique <sup>b</sup>	Study location	MAP (mm) <sup>c</sup>	MAT (°C)
Borhan et al. (2011a)	dynamic chamber	GC	central Texas	501	14.8
Leytem et al. (2011)	sample tower, bLS	PAS	southern Idaho, USA	254	9.4
Borhan et al. (2011b)	dynamic chamber	GC	central TX	501	14.8
Bjorneberg et al. (2009)	open path FTIR, bLS	FTIR	southern Idaho, USA	254	9.4
Craggs et al. (2008)	large static chamber	IRGA	Hamilton, NZ	1160	15.1
Safley and Westerman (1988)	large static chamber	gas concentration tubes	Raleigh, NC, USA	1059	15.1
Safley and Westerman (1992)	large static chamber	gas concentration tubes	Raleigh, NC, USA	1059	15.1
Todd et al. (2011)	open path laser	TDLAS	Curry County, NM	424	23.6
Husted (1994)	dynamic chamber	GC	Askov, Denmark	862	7.7
Sneath et al. (2006) <sup>d</sup>	SF <sub>6</sub> tracer	GC	Brittany, France/southwest England	730/415	11.7/5.5
Hensen et al. (2006)	Gaussian plume model	TDLAS	Wageningen, the Netherlands	820	9.4
Kaharabata et al. (1998)	SF <sub>6</sub> tracer	precision gas detector	St-Anne-de-Bellevue, Quebec, Canada	1036	6.3
VanderZaag et al. (2011)	IHF	TDLAS	Bright, Ontario, Canada	858	6.9
Ahn et al. (2011)	large dynamic chamber	PAS	Beltsville, MD, USA	1058	5
Amon et al. (2006)	large dynamic chamber	FTIR	Gross Enzersdorf, Austria	547	9.8
Brown et al. (2002)	sample mast, IHF	TDLAS	Ottawa, Ontario, Canada	800	6.7
Gupta et al. (2007)	static chamber	GC	New Delhi, India	706	25
Osada et al. (2001)	large dynamic chamber	GC	Denmark	600	7.8
Sommer et al. (2004)	sample masts, IHF	GC	Lethbridge, Alberta, Canada	415	5.5
Kaharabata et al. (2000)	SF <sub>6</sub> tracer	GC	Ottawa, Ontario, Canada	862	5.9
Ellis et al. (2001)	static chamber	PAS-IRGA	Shropshire, England	747	9.2
Gao et al. (2011)	open path laser, bLS	TDLAS	Baoding, Hebei, China	541	12.3
Misselbrook et al. (2001)	static chamber	GC or PAS	Shropshire, England	673	9.2
Adviento-Borbe et al. (2010)	static chamber	PAS	State College, PA, USA		9.3
Kinsman et al. (1995)	exhaust fan sample	IRGA	Ottawa, Canada		5.9
Marik and Levin (1996)	ventillation rate, SF <sub>6</sub> tracer	GC	Germany		10.6

<b>Study</b>	<b>Measurement technique<sup>a</sup></b>	<b>Gas analysis technique<sup>b</sup></b>	<b>Study location</b>	<b>MAP (mm)<sup>c</sup></b>	<b>MAT (°C)</b>
Ngwabie et al. (2009)	CO <sub>2</sub> tracer	PAS	southern Sweden		7.8
Samer et al. (2012)	ventilation rate, <sup>85</sup> Kr tracer	multigas monitor, radiation counter	NE Germany		8.4
Snell et al. (2003)	SF <sub>6</sub> tracer	PAS	NW Germany		8.6
van Vliet et al. (2004)	static chamber	GC	Agassiz, British Columbia, Canada		10
Zhang et al. (2005)	SF <sub>6</sub> or CO <sub>2</sub> tracer	PAS	Denmark		7.7
McGinn and Beauchemin (2012)	open path laser, inverse dispersion	TDLAS	Lethbridge, Alberta		5.5

<sup>a</sup> FTIR = Fourier transform infrared spectroscopy, bLS = backwards Langrangian stochastic dispersion technique, IHF = integrated horizontal flux

<sup>b</sup> PAS = photoacoustic field gas monitor, GC = gas chromatograph, TDLAS = tunable diode laser absorption spectrometer, IRGA = infrared gas analyzer

<sup>c</sup> MAP not reported for studies which measured only barns

<sup>d</sup> location and climate data reported for slurry tank and manure pile, respectively

**Table A2. Anaerobic lagoon field measurements.**

<b>Paper</b>	<b>What studied</b>	<b>Sample period</b>	<b>Sampling T (°C)</b>	<b>Number of cows</b>	<b>Amount of slurry (m<sup>3</sup>)</b>	<b>Surface area (ha)</b>	<b>Depth of pond (m)</b>	<b>N<sub>2</sub>O (kg/ha/d)</b>	<b>N<sub>2</sub>O (g/hd/d)</b>	<b>CH<sub>4</sub> (kg/ha/d)</b>	<b>CH<sub>4</sub> (g/hd/d)</b>
Bjorneberg et al. (2009)	anaerobic lagoon	January	4	780		0.8				16	16
Bjorneberg et al. (2009)	anaerobic lagoon	March	5	780		0.8				15	15
Bjorneberg et al. (2009)	anaerobic lagoon	June	22	780		0.8				41	42
Bjorneberg et al. (2009)	anaerobic lagoon	September	12	780		0.8				25	26
Borhan et al. (2011a)	anaerobic lagoon (primary)	summer	26	3200		2.32		0.04	0.027	919	666
Borhan et al. (2011a)	anaerobic lagoon (secondary)	summer	26	3200		4.01		0.03	0.034	112	141
Borhan et al. (2011b)	primary anaerobic lagoon	summer	34	500		0.05		0.10	0.01	840	85
Borhan et al. (2011b)	primary anaerobic lagoon	winter	11	500		0.75		0.31	0.47	21	31
Borhan et al. (2011b)	secondary anaerobic lagoon	winter	13	500		0.70		0.35	0.49	7	10
Craggs et al. (2008)	anaerobic lagoon	2006	15	700	4600	0.18	5.0			186	47
Leytem et al. (2011)	anaerobic lagoon	Apr 08	14	13000		10		8.5	6.5	194	150
Leytem et al. (2011)	anaerobic lagoon	May 08	15	13000		10		7.0	5.4	961	739
Leytem et al. (2011)	anaerobic lagoon	Aug 08	21	13000		10		1.2	0.9	653	502
Leytem et al. (2011)	anaerobic lagoon	Oct 08	15	13000		10		2.9	2.2	2307	1774
Safley and Westerman (1988)	anaerobic lagoon	August 85	24	130	11770	0.5	2.6			401	1542
Safley and Westerman (1988)	anaerobic lagoon	April-Sept 86	23	130	11770	0.5	2.6			2005	7710
Safley and Westerman (1992)	anaerobic lagoon	summer 87	26	200	11770	0.5	2.6			1053	2631
Todd et al. (2011)	anaerobic lagoon	August	24	3500	27000	1.80				402	207

*Note:* Blanks indicate data not provided or measured.

**Table A3. Slurry tank field measurements.**

<b>Paper</b>	<b>Sample period</b>	<b>Sampling T (°C)</b>	<b>Number of cows</b>	<b>Manure N content (g/kg FM)</b>	<b>%DM in fresh manure</b>	<b>Volatile solids (g/kg)</b>	<b>Amount of slurry (m<sup>3</sup>)</b>	<b>Surface area (ha)</b>	<b>Depth (m)</b>
Borhan et al. (2011a)	summer	26	3200					0.127	
Borhan et al. (2011b)	summer	31.2	500					0.089	
Borhan et al. (2011b)	winter	14.6	500					0.089	
Hensen et al. (2006)	Feb 03	9.4					1200		
Husted (1994)	Oct 91	9.75	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Nov 91	5.5	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Dec 91	3	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Jan 92	-4.125	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Feb 92	1.75	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Mar 92	5.625	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Apr 92	6.125	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	May 92	14.125	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	June 92	20.5	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	July 92	14.875	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Aug 92	16.25	120	3.56	6.48	47.5	1200	0.03	4
Husted (1994)	Sept 92	14.5	120	3.56	6.48	47.5	1200	0.03	4
Kaharabata et al. (1998)	June-Nov 95	6.3	110		7.5		2260	0.045	5
Sneath et al. (2006)	Apr 02	11	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	May 02	17	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	June 02	21	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	July 02	18	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	Aug 02	20	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	Sept 02	19	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	Oct 02	17	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	Nov 02	14	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	early Dec 02	11	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	late Dec 02	12	122.4	2.75	5.1	39	2150	0.080	2.7
Sneath et al. (2006)	Feb 02	13	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Mar 02	12	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Apr 02	11	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	May 02	17	126	2.75	6.6	44	1500	0.045	3.3

<b>Paper</b>	<b>Sample period</b>	<b>Sampling T (°C)</b>	<b>Number of cows</b>	<b>Manure N content (g/kg FM)</b>	<b>%DM in fresh manure</b>	<b>Volatile solids (g/kg)</b>	<b>Amount of slurry (m<sup>3</sup>)</b>	<b>Surface area (ha)</b>	<b>Depth (m)</b>
Sneath et al. (2006)	June 02	21	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	July 02	18	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Aug 02	20	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Sept 02	19	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Oct 02	17	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	Nov 02	14	126	2.75	6.6	44	1500	0.045	3.3
Sneath et al. (2006)	early Dec 02	11	126	2.75	6.6	44	1500	0.045	3.3
VanderZaag et al. (2011)	Jan 03	-10.6	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	Feb 03	-6.9	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	Mar 03	1.9	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	Apr 03	6.8	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	May 03	11	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	Jun 03	15.2	100	4.5	7.1		1429.2	0.040	3.6
VanderZaag et al. (2011)	July 03	20.1	100	4.5	7.1		1429.2	0.040	3.6

*Note:* Blanks indicate data not provided or not measured.

<b>Paper</b>	<b>Crust</b>	<b>Aerated</b>	<b>CO<sub>2</sub> (kg/ha/d)</b>	<b>N<sub>2</sub>O (kg/ha/d)</b>	<b>CH<sub>4</sub> (kg/ha/d)</b>	<b>CO<sub>2</sub> (g/hd/d)</b>	<b>N<sub>2</sub>O (g/hd/d)</b>	<b>CH<sub>4</sub> (g/hd/d)</b>	<b>N<sub>2</sub>O (g/m<sup>3</sup>/y slurry)</b>	<b>CH<sub>4</sub> (g/m<sup>3</sup> slurry/y)</b>
Borhan et al. (2011a)		n	1795	0.10	359	71	0.004	14		
Borhan et al. (2011b)		n	1261	0.17	930	225	0.030	166		
Borhan et al. (2011b)		n	504	0.28	26	90	0.050	5		
Hensen et al. (2006)									5.1	4015
Husted (1994)	n	y			369	-		92		3369
Husted (1994)	n	y			164	-		41		1497
Husted (1994)	n	y			287	-		72		2621
Husted (1994)	n	y			0	-		0		0
Husted (1994)	y	y			21	-		5		187
Husted (1994)	y	y			31	-		8		281
Husted (1994)	y	y			51	-		13		468
Husted (1994)	n	y			390	-		97		3556
Husted (1994)	y	y			359	-		90		3276
Husted (1994)	n	y			708	-		177		6458
Husted (1994)	n	y			1436	-		359		13103
Husted (1994)	n	y			164	-		41		1497
Kaharabata et al. (1998)	y	some			2027	-		833		14800
Sneath et al. (2006)	y	0		2.67	991	-	1.757	651	36.5	13526
Sneath et al. (2006)	y	0		2.29	676	-	1.506	444	31.3	9232
Sneath et al. (2006)	y	0		10.31	959	-	6.775	630	140.8	13097
Sneath et al. (2006)	y	0		4.20	1179	-	2.760	775	57.4	16103
Sneath et al. (2006)	y	0		3.44	1368	-	2.258	899	46.9	18679
Sneath et al. (2006)	y	0		0.38	991	-	0.251	651	5.2	13526
Sneath et al. (2006)	y	0		9.17	1368	-	6.022	899	125.1	18679
Sneath et al. (2006)	y	0		18.71	865	-	12.296	568	255.5	11809
Sneath et al. (2006)	y	0		11.84	629	-	7.779	413	161.6	8588
Sneath et al. (2006)	y	0		0.38	1022	-	0.251	672	5.2	13956
Sneath et al. (2006)	y	0		0.00	410	-	0	147	0.0	4509
Sneath et al. (2006)	y	0		0.95	527	-	0.340	189	10.4	5797
Sneath et al. (2006)	y	0		0.00	546	-	0.000	196	0.0	6012
Sneath et al. (2006)	y	0		0.47	780	-	0.170	280	5.2	8588

<b>Paper</b>	<b>Crust</b>	<b>Aerated</b>	<b>CO<sub>2</sub> (kg/ha/d)</b>	<b>N<sub>2</sub>O (kg/ha/d)</b>	<b>CH<sub>4</sub> (kg/ha/d)</b>	<b>CO<sub>2</sub> (g/hd/d)</b>	<b>N<sub>2</sub>O (g/hd/d)</b>	<b>CH<sub>4</sub> (g/hd/d)</b>	<b>N<sub>2</sub>O (g/m<sup>3</sup>/y slurry)</b>	<b>CH<sub>4</sub> (g/m<sup>3</sup> slurry/y)</b>
Sneath et al. (2006)	y	0		0.95	702	-	0.340	252	10.4	7729
Sneath et al. (2006)	y	0		5.21	1287	-	1.871	462	57.4	14171
Sneath et al. (2006)	y	0		2.37	1443	-	0.850	518	26.1	15888
Sneath et al. (2006)	y	0		0.95	819	-	0.340	294	10.4	9018
Sneath et al. (2006)	y	0		0.47	1248	-	0.170	448	5.2	13741
Sneath et al. (2006)	y	0		0.47	663	-	0.170	238	5.2	7300
Sneath et al. (2006)	y	0		0.47	488	-	0.170	175	5.2	5368
VanderZaag et al. (2011)	y	some			221	-		88		2243
VanderZaag et al. (2011)	y	some			264	-		105		2681
VanderZaag et al. (2011)	y	some			338	-		134		3425
VanderZaag et al. (2011)	y	some			377	-		150		3819
VanderZaag et al. (2011)	y	some			420	-		167		4257
VanderZaag et al. (2011)	y	some			96	-		38		972
VanderZaag et al. (2011)	y	some			1323	-		525		13412

*Note:* Blanks indicate data not provided or not measured.



**Table A4. Solid system (manure pile and manure compost area) field measurements.**

<b>Paper</b>	<b>GHG measurement technique</b>	<b>What studied</b>	<b>Turning protocol</b>	<b>Nanure additions</b>	<b>Sample period</b>	<b>T during sampling (°C)</b>
Ahn et al. (2011)	large dynamic chamber	solid manure pile (static)	none	0	na	5
Ahn et al. (2011)	large dynamic chamber	solid manure pile (mixed)	4x/80 days	0	na	5
Amon et al. (2006)	large dynamic chamber	solid manure pile	7x	0	na	17
Borhan et al. (2011a)	dynamic chamber	compost pile		2x week	summer	26
Brown et al. (2002)	IHF	solid manure pile	none	daily	June	18.5
Gupta et al. (2007)	static chamber	solid manure pile	na	daily	Nov 02	20.3
Gupta et al. (2007)	static chamber	solid manure pile	na	daily	Jun 03	33.8
Gupta et al. (2007)	static chamber	solid manure pile	na	daily	July 03	31
Husted (1994)	dynamic chamber	solid manure pile	none	daily	Oct 91-Sep 92	7.7
Leytem et al. (2011)	sample tower	composting area	na	daily	year avg	
Leytem et al. (2011)	sample tower	composting area	na	daily	Mar 08	13
Leytem et al. (2011)	sample tower	composting area	na	daily	Apr 08	8.1
Leytem et al. (2011)	sample tower	composting area	na	daily	May 08	9.8
Leytem et al. (2011)	sample tower	composting area	na	daily	June 08	23.6
Leytem et al. (2011)	sample tower	composting area	na	daily	Aug 08	24.4
Leytem et al. (2011)	sample tower	composting area	na	daily	Sept 08	15
Osada et al. (2001)	large dynamic chamber	manure and barley straw (deep bedding)	3 weeks	0	Jul-Aug 97	18
Sneath et al. (2006)	SF6 tracer	solid manure pile in a naturally ventilated building	1x/month	0	June-Aug 02	14
Sommer et al. (2004)	IHF	solid manure pile	none	0	Sept 03	11

*Note:* Blanks indicate data not provided or not measured.

<b>Paper</b>	<b>Number of cows</b>	<b>Manure in pile (kg)</b>	<b>Manure in pile (m<sup>3</sup>)</b>	<b>Manure N (g/kg DM)</b>	<b>Manure C (g/kg DM)</b>	<b>DM (%)</b>	<b>VS (g/100g DM)</b>	<b>Surface area (m<sup>2</sup>)</b>
Ahn et al. (2011)		1200		0.18	4.80	24.0	93	1.58
Ahn et al. (2011)		1200		0.18	4.80	24.0	93	1.58
Amon et al. (2006)		-	10	1.07	16.89	20.9	16.04	4.91
Borhan et al. (2011a)	6950							30674
Brown et al. (2002)	310		800	0.11		19.0		400
Gupta et al. (2007)								
Gupta et al. (2007)								
Gupta et al. (2007)								
Husted (1994)	32			0.90		18.7	76.19	
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Leytem et al. (2011)	13000							100000
Osada et al. (2001)		1080	4.60	3.92	65.02	40.9	34.98	7.03
Sneath et al. (2006)	150	200000		0.84		26.7		
Sommer et al. (2004)		2700	3.28	0.37	6.25	44.8		6.41

*Note:* Blanks indicate data not provided or not measured.

Paper	N <sub>2</sub> O (kg/ha/d)	N <sub>2</sub> O (g/m <sup>3</sup> /y)	N <sub>2</sub> O (mg/kg/d)	N <sub>2</sub> O (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/m <sup>3</sup> /y)	CH <sub>4</sub> (mg/kg/d)	CH <sub>4</sub> (g/hd/d)
Ahn et al. (2011)	1.32		174		94.74		0.01	
Ahn et al. (2011)	4.59		606		136.84		0.02	
Amon et al. (2006)	3.36	13.20			130.02	511		
Borhan et al. (2011a)	0.13			0.06	0.09			4.00E-02
Brown et al. (2002)	13.20	240.90		1.70				
Gupta et al. (2007)			31.29	6.26E-05			4.29	8.58E-03
Gupta et al. (2007)			72.11	1.44E-04			4.84	9.68E-03
Gupta et al. (2007)			6.39	1.28E-05			12.92	2.58E-02
Husted (1994)						2154		1.45E-05
Leytem et al. (2011)	9.00	-			135.00			103.85
Leytem et al. (2011)	1.20			0.92	150.60			115.85
Leytem et al. (2011)	7.90			6.08	166.00			127.69
Leytem et al. (2011)	12.10			9.31	47.60			36.62
Leytem et al. (2011)	26.70			20.54	352.20			270.92
Leytem et al. (2011)	4.30			3.31	70.30			54.08
Leytem et al. (2011)	1.90			1.46	25.80			19.85
Osada et al. (2001)	3.50	195.37	2280		4.45		2.90	
Sneath et al. (2006)		471.48				8322		
Sommer et al. (2004)	17.16	1223.31	4074		79.56	5672	18.89	

*Note:* Blanks indicate data not provided or not measured.

**Table A5. Corral field measurements.**

Paper	Type of corral	Soil texture	Scraping frequency	Includes cows	Season/months	Sampling T (°C)	Lot conditions
Bjorneberg et al. (2009)	open lot	sandy loam	daily	y	Jan 05	-2.95	na
Bjorneberg et al. (2009)	open lot	sandy loam	daily	y	Mar 05	6.3	na
Bjorneberg et al. (2009)	open lot	sandy loam	daily	y	Aug 05	25.65	na
Bjorneberg et al. (2009)	open lot	sandy loam	daily	y	Sept 05	14.05	na
Borhan et al. (2011a)	loafing pen	loam to clay loam	na	n	summer	26	dry
Borhan et al. (2011a)	open lot	loam to clay loam	2x/wk	n	summer	26	dry
Borhan et al. (2011b)	loafing pen	loam to clay loam	na	n	summer	36.1	dry
Borhan et al. (2011b)	loafing pen	loam to clay loam	na	n	winter	10.1	wetter
Kaharabata et al. (2000)	open lot	sand	na	y	Aug 96	23	na
Leytem et al. (2011)	open lot	sandy loam	na	y	Mar 08	8.35	partially frozen, wet
Leytem et al. (2011)	open lot	sandy loam	na	y	Apr 08	13.68	saturated
Leytem et al. (2011)	open lot	sandy loam	na	y	May 08	20.6	dry
Leytem et al. (2011)	open lot	sandy loam	na	y	Jun 08	22.24	dry
Leytem et al. (2011)	open lot	sandy loam	na	y	Aug 08	23.49	dry
Leytem et al. (2011)	open lot	sandy loam	na	y	Sept 08	11.68	dry
Leytem et al. (2011)	open lot	sandy loam	na	y	Oct 08	12.43	part wet
Leytem et al. (2011)	open lot	sandy loam	na	y	Nov 08	8.43	part wet
<i>Leytem et al. (2011)*</i>	<i>open lot</i>	<i>sandy loam</i>	<i>na</i>	<i>y</i>	<i>year avg</i>	-	

\*year average was not included in the mean in Table 2

Note: na =not available, nm = not measured

Paper	Number of cows	Milk	Heifer/dry	Surface area (ha)	N <sub>2</sub> O		<i>as measured</i>		<i>no enteric emissions</i>	
					(kg/ha/d)	(g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)
Bjorneberg et al. (2009)	780	700	80	4.5	nm	nm	51	294	3.7	22
Bjorneberg et al. (2009)	780	700	80	4.5	nm	nm	83	479	36	206
Bjorneberg et al. (2009)	780	700	80	4.5	nm	nm	26	150	-21	-123
Bjorneberg et al. (2009)	780	700	80	4.5	nm	nm	26	150	-21	-123
Borhan et al. (2011a)	3200	3200	0	3.61	0.53	0.60	2.7	3.1	2.7	3.1
Borhan et al. (2011a)	3750	0	3750	12.4	0.73	2.41	3.3	11.0	3.3	11
Borhan et al. (2011b)	500	500	0	2.26	1.66	7.50	4.7	21.5	4.7	22
Borhan et al. (2011b)	500	500	0	2.26	0.49	2.20	1.1	5.1	1.1	5.1
Kaharabata et al. (2000)	147	0	147	0.36	nm	nm	20	0.1	-90	-222
Leytem et al. (2011)	10800	10800	0	60	0.83	4.63	96	533	47	261
Leytem et al. (2011)	10800	10800	0	60	3.08	17.13	167	926	118	653
Leytem et al. (2011)	10800	10800	0	60	5.98	33.24	113	626	64	353
Leytem et al. (2011)	10800	10800	0	60	2.33	12.96	63	352	14	79
Leytem et al. (2011)	10800	10800	0	60	1.08	6.02	85	470	36	197
Leytem et al. (2011)	10800	10800	0	60	1.03	5.74	12	69	-37	-203
Leytem et al. (2011)	10800	10800	0	60	0.03	0.19	90	501	41	229
Leytem et al. (2011)	10800	10800	0	60	0.35	1.94	67	372	18	99
<i>Leytem et al. (2011)*</i>	<i>10800</i>	<i>10800</i>	<i>0</i>	<i>60</i>	<i>1.83</i>	<i>10.19</i>	<i>87</i>	<i>485</i>	<i>38</i>	<i>213</i>

\*year average was not included in the mean in Table 2

Note: na =not available, nm = not measured

**Table A6. Emissions measurements from surfaces.**

<b>Paper</b>	<b>What measured</b>	<b>Sample period</b>	<b>Manure management</b>	<b>Includes cows</b>	<b>Sampling T (°C)</b>
<i>hardstandings</i>					
Gao et al. (2011)	brick-floored pen	Aug-Sept 09	scraped every 1-3 weeks	y	12.3
Gao et al. (2011)	brick-floored pen	Feb 10	scraped 1x/month	y	12.3
Ellis et al. (2001)	feeding/loafing area	summer	na	n	9.2
Ellis et al. (2001)	concrete collecting yard	summer	scraped daily	n	9.2
Ellis et al. (2001)	concrete collecting yard	winter	scraped daily	n	9.2
Misselbrook et al. (2001)	concrete collecting yard	Sept 97-Jan 00	scraped daily	n	9.2
Misselbrook et al. (2001)	concrete collecting yard	Sept 97-Jan 00	scraped daily	n	9.4
Misselbrook et al. (2001)	concrete collecting yard	Sept 97-Jan 00	scraped daily	n	10.7
Misselbrook et al. (2001)	concrete feeding/loafing yard	Sept 97-Jan 00	scraped daily	n	10.7
Misselbrook et al. (2001)	concrete feeding/loafing yard	Sept 97-Jan 00	infrequent scraping	n	9.2
<i>barn floors</i>					
Borhan et al. (2011a)	freestall bedding	summer	na	n	26
Borhan et al. (2011a)	manure alley	summer	alley flushed every 2 h	n	26
Borhan et al. (2011a)	freestall bedding	summer	na	n	26.2
Borhan et al. (2011a)	freestall bedding	winter	na	n	8.4
Borhan et al. (2011b)	manure alley	summer	alley flushed 1x/d	n	23.8
Borhan et al. (2011b)	manure alley	winter	alley flushed 1x/d	n	8.4
van Vliet et al. (2004)	manure alley next to sand deep bedding	Jun 2002	beds raked into alley, alleys scraped every 4 h	n	na
van Vliet et al. (2004)	manure alley next to sawdust deep bedding	Jun 2002	beds raked into alley, alleys scraped every 4 h	n	na
van Vliet et al. (2004)	manure alley next to sand deep bedding	Mar 2003	beds raked into alley, alleys scraped every 4 h	n	na
van Vliet et al. (2004)	manure alley next to sawdust deep bedding	Mar 2003	beds raked into alley, alleys scraped every 4 h	n	na
Adviento-Borbe et al. (2010)	barn floor, alfalfa/corn feed	Feb-May 05	scraped 2x/day	n	7.21
Adviento-Borbe et al. (2010)	barn floor, grass/corn	Feb-May 05	scraped 2x/day	n	7.75
Adviento-Borbe et al. (2010)	barn floor, alfalfa/corn feed	June-Oct 05	scraped 2x/day	n	21
Adviento-Borbe et al. (2010)	barn floor, corn/corn feed	June-Oct 05	scraped 2x/day	n	21.5

*Note:* Sampling temperature for the hardstandings is MAT and for the barn floors is the indoor barn temperature.

<b>Paper</b>	<b>Number of cows</b>	<b>Surface area (m2)</b>	<b>CO<sub>2</sub> (kg/ha/d)</b>	<b>N<sub>2</sub>O (kg/ha/d)</b>	<b>CH<sub>4</sub> (kg/ha/d)</b>	<b>CO<sub>2</sub> (g/hd/d)</b>	<b>N<sub>2</sub>O (g/hd/d)</b>	<b>CH<sub>4</sub> (g/hd/d)</b>
<i>hardstandings</i>								
Gao et al. (2011)	682	20300	nm	nm	23	nm	nm	67
Gao et al. (2011)	590	20300	nm	nm	11	nm	nm	37
Ellis et al. (2001)	135	200	nm	0.0135	48	nm	0.0020	7.2
Ellis et al. (2001)	135	150	nm	0.0049	18	nm	0.0005	2.0
Ellis et al. (2001)	135	150	nm	0.0025	59	nm	0.0003	6.6
Misselbrook et al. (2001)	135	150	nm	0.0113	0.166	nm	0.0013	0.018
Misselbrook et al. (2001)	240	200	nm	0.0022	0.079	nm	0.0002	0.007
Misselbrook et al. (2001)	85	82	nm	0.0045	0.086	nm	0.0004	0.008
Misselbrook et al. (2001)	85	156	nm	0.0095	0.098	nm	0.0017	0.018
Misselbrook et al. (2001)	135	150	nm	0.0296	0.048	nm	0.0033	0.005
<i>barn floors</i>								
Borhan et al. (2011a)	3200	8451	261	0.765	0.49	69	0.202	0.13
Borhan et al. (2011a)	3200	12527	227	0.066	0.87	89	0.026	0.34
Borhan et al. (2011a)	500	1524	846	0.984	2.17	258	0.300	0.66
Borhan et al. (2011a)	500	1524	341	0.361	0.89	104	0.110	0.27
Borhan et al. (2011b)	500	1980	449	0.076	2.63	178	0.030	1.04
Borhan et al. (2011b)	500	1980	624	0.404	1.46	247	0.160	0.58
van Vliet et al. (2004)	12	115.2	170	0.017	0.48	164	0.016	0.46
van Vliet et al. (2004)	12	115.2	137	0.012	0.41	131	0.012	0.39
van Vliet et al. (2004)	12	115.2	191	0.029	1.44	184	0.028	1.38
van Vliet et al. (2004)	12	115.2	257	0.007	0.50	247	0.007	0.48
Adviento-Borbe et al. (2010)	60	257	1134	0.008	11.83	486	0.004	5.07
Adviento-Borbe et al. (2010)	60	257	1112	0.008	13.76	477	0.004	5.90
Adviento-Borbe et al. (2010)	60	257	1486	0.011	19.81	637	0.005	8.50
Adviento-Borbe et al. (2010)	60	257	1596	0.011	28.07	684	0.005	12.04

**Table A7. Field measurements of whole barns.**

Paper	Flooring	Manure management	Subfloor manure storage	Deep bedding	Sample period	Includes cows	Indoor sampling T (°C)	Ventilation rate (m <sup>3</sup> /hd/d)
Kinsman et al. (1995)	paved	scraped gutter to subfloor tank	y	n	June-Nov 93	n		
Marik and Levin (1996)	slatted	gutters drain to outside	n	na	27-Mar-92	y		5581
Ngwabie et al. (2009)	slatted	gutters scraped 2x/day	na	some	Dec 20-31, 2006	y	13	301
Ngwabie et al. (2009)	slatted	gutters scraped 2x/day	na	some	Jan 11-31, 2007	y	12	317
Ngwabie et al. (2009)	slatted	gutters scraped 2x/day	na	some	Feb 2007	y	10	300
Ngwabie et al. (2009)	slatted	gutters scraped 2x/day	na	some	Mar 2007	y	14	318
Ngwabie et al. (2009)	slatted	gutters scraped 2x/day	na	some	May 22-29, 2007	s	18	481
Samer et al. (2012)	concrete	continuously scraped alleys	na	n	summer	y	21	108306
Samer et al. (2012)	concrete	continuously scraped alleys	na	n	winter	y	3	70876
Snell et al. (2003)	slatted, concrete	scraped gutters	na	n	Jan 1998	y		302
Snell et al. (2003)	asphalt	alleys scraped to outside	n	n	Feb 1997	y		354
Snell et al. (2003)	solid	alleys scraped to outside	na	n	Sep 1998	y		710
Snell et al. (2003)	walkways slatted	na	na	n	Apr 1998	y		351
Zhang et al. (2005)	concrete	delta scraper	na	n	summer	y	21.7	
Zhang et al. (2005)	concrete	delta scraper	na	n	winter	y	7	
Zhang et al. (2005)	asphalt	liquid drain and solid scraper	na	n	summer	y	17.4	
Zhang et al. (2005)	asphalt	liquid drain and solid scraper	na	n	winter	y	6.2	
Zhang et al. (2005)	grooved concrete	liquid drain and solid scraper	na	n	summer	y	19.9	
Zhang et al. (2005)	grooved concrete	liquid drain and solid scraper	na	n	winter	y	6.1	
Zhang et al. (2005)	grooved concrete	liquid drain and solid scraper	na	n	winter	y	6.4	
Zhang et al. (2005)	concrete	liquid drain and solid scraper	na	n	summer	y	15.7	
Zhang et al. (2005)	concrete	liquid drain and solid scraper	na	n	winter	y	3.6	
Zhang et al. (2005)	slatted	scraped gutters	na	n	winter	y	9.6	
Zhang et al. (2005)	slatted	scraped gutters	na	n	summer	y	19	
Zhang et al. (2005)	slatted	back flushed gutters	na	n	summer	y	22.4	
Zhang et al. (2005)	slatted	back flushed gutters	na	n	winter	y	8.3	
Zhang et al. (2005)	slatted	circulation (no acid)	na	n	summer	y	13.2	



<b>Paper</b>	<b>Flooring</b>	<b>Manure management</b>	<b>Subfloor manure storage</b>	<b>Deep bedding</b>	<b>Sample period</b>	<b>Includes cows</b>	<b>Indoor sampling T (°C)</b>	<b>Ventilation rate (m<sup>3</sup>/hd/d)</b>
Zhang et al. (2005)	slatted	circulation (no acid)	na	n	winter	y	2.3	
Zhang et al. (2005)	slatted	circulation (with acid)	na	n	summer	y	16.8	
Zhang et al. (2005)	slatted	circulation (with acid)	na	n	summer	y	16.8	
Zhang et al. (2005)	slatted	circulation (with acid)	na	n	summer	y	14.7	
Zhang et al. (2005)	slatted	circulation (with acid)	na	n	winter	y	6.2	
Zhang et al. (2005)	slatted	slats scraped, circulation	na	n	summer	y	19.5	
Zhang et al. (2005)	slatted	slats scraped, circulation	na	n	winter	y	5.8	
Zhang et al. (2005)	slatted	circulation (no additive)	na	n	summer	y	16.6	
Zhang et al. (2005)	slatted	circulation (no additive)	na	n	summer	y	17.3	
Zhang et al. (2005)	slatted	circulation (no additive)	na	n	winter	y	6.3	
Zhang et al. (2005)	slatted	circulation (with additive)	na	n	winter	y	8.9	

*Note: Enclosed tie stall barns were measured by Marik and Levin (1996) and Kinsman et al. (1995). The other studies measured enclosed freestall barns.*

Paper	Number of cows	Average cow mass (kg)	Surface area (m <sup>2</sup> )	<i>As measured</i>				<i>No enteric emissions</i>	
				N <sub>2</sub> O (kg/ha/d)	N <sub>2</sub> O (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)
Kinsman et al. (1995)	118	602	832			36	25	36	25
Marik and Levin (1996)	43		512			247	294	17	21
Ngwabie et al. (2009)	195	600	2425			276	343	55	69
Ngwabie et al. (2009)	184	600	2425			284	374	76	100
Ngwabie et al. (2009)	189	600	2425			260	334	47	60
Ngwabie et al. (2009)	170	600	2425			228	325	36	51
Ngwabie et al. (2009)	164	600	2425			175	259	67	99
Samer et al. (2012)	348	669	3288	62.8	59.3	1177	1111	887	837
Samer et al. (2012)	336	671	3288	62.7	61.4	615	602	335	328
Snell et al. (2003)	105	579	858			552	451	216	177
Snell et al. (2003)	51	660	856			210	352	47	78
Snell et al. (2003)	118	383	1080			655	599	355	325
Snell et al. (2003)	86	656	554			709	456	283	182
Zhang et al. (2005)	165	600	2276	0.23	0.32	463	638	264	364
Zhang et al. (2005)	181	600	2276	0.64	0.8	310	390	93	116
Zhang et al. (2005)	129	600	2293	0.00	0	220	391	66	117
Zhang et al. (2005)	122	600	2293	3.16	5.94	277	520	131	246
Zhang et al. (2005)	282	600	2023	1.95	1.4	425	305	44	31
Zhang et al. (2005)	170	600	2023	0.22	0.26	258	307	28	33
Zhang et al. (2005)	180	600	2023	0.37	0.42	308	346	64	72
Zhang et al. (2005)	105	600	2421	1.53	3.52	172	397	53	123
Zhang et al. (2005)	115	600	2421	0.00	0	175	369	45	95
Zhang et al. (2005)	152	600	2965	1.23	2.4	200	390	60	116
Zhang et al. (2005)	148	600	2965	0.64	1.28	271	542	134	268
Zhang et al. (2005)	159	600	1536	0.16	0.15	533	515	249	241
Zhang et al. (2005)	170	600	1536	0.00	0	352	318	49	44
Zhang et al. (2005)	192	450	1221	8.68	5.52	934	594	503	320
Zhang et al. (2005)	192	450	1221	12.83	8.16	572	364	141	90
Zhang et al. (2005)	192	450	1221	3.88	2.47	558	355	127	81
Zhang et al. (2005)	192	450	1221	2.86	1.82	527	335	97	61
Zhang et al. (2005)	192	450	1221	0.19	0.12	540	343	109	69
Zhang et al. (2005)	192	450	1221	0.94	0.6	472	300	41	26

Paper	Number of cows	Average cow mass (kg)	Surface area (m <sup>2</sup> )	<i>As measured</i>		<i>No enteric emissions</i>			
				N <sub>2</sub> O (kg/ha/d)	N <sub>2</sub> O (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)	CH <sub>4</sub> (kg/ha/d)	CH <sub>4</sub> (g/hd/d)
Zhang et al. (2005)	115	600	1582	1.40	1.92	343	472	144	198
Zhang et al. (2005)	121	600	1582	1.61	2.1	325	425	115	151
Zhang et al. (2005)	166	600	2004	3.80	4.59	414	500	187	226
Zhang et al. (2005)	159	600	2004	4.70	5.92	352	443	134	169
Zhang et al. (2005)	163	600	2004	4.03	4.96	363	446	140	172
Zhang et al. (2005)	163	600	2004	3.64	4.48	371	456	148	182

**Table A8. Field measurements of emissions plumes from whole dairies.**

Source	Dairy ID	Study location	Sampling season/month	Sampling season mean T (°C)	MAT (°C)	Number of cows adding manure stores
Hensen et al. (2005,2006)	A1	Hardinxveld, the Netherlands	Mar-02	5.9	10	65
Hensen et al. (2005,2006)	A2	Hardinxveld, the Netherlands	Mar-02	5.9	10	65
Hensen et al. (2005,2006)	A3	Hardinxveld, the Netherlands	Mar-02	5.9	10	40
Hensen et al. (2005,2006)	A4	Hardinxveld, the Netherlands	Mar-02	5.9	10	93
Hensen et al. (2005,2006)	A5	Hardinxveld, the Netherlands	Mar-02	5.9	10	85
Hensen et al. (2005,2006)	B1	North Holland, the Netherlands	Jun-03	17	10	155
Hensen et al. (2005,2006)	B1	North Holland, the Netherlands	Jan-04	2.9	10	143
Hensen et al. (2005,2006)	B2	North Holland, the Netherlands	Jun-03	17	10	115
Hensen et al. (2005,2006)	B3	North Holland, the Netherlands	Jun-03	17	10	29
Hensen et al. (2005,2006)	B4	North Holland, the Netherlands	Jun-03	17	10	196
Hensen et al. (2005,2006)	B5	North Holland, the Netherlands	Jun-03	17	10	170
Hensen et al. (2005,2006)	B6	North Holland, the Netherlands	Jun-03	17	10	18
Hensen et al. (2005,2006)	C1	Brabant, the Netherlands	Jun-03	18	10	600
Hensen et al. (2005,2006)	C2	Brabant, the Netherlands	Jun-03	18	10	160
Hensen et al. (2005,2006)	C3	Brabant, the Netherlands	Jun-03	18	10	93
McGinn and Beauchemin (2012)	dairy 1	Lethbridge, Alberta	Nov-04	-0.7	5.5	302
McGinn and Beauchemin (2012)	dairy 2	Lethbridge, Alberta	May-05	11.1	5.5	207
McGinn and Beauchemin (2012)	dairy 3	Lethbridge, Alberta	Jul-05	18.6	5.5	351

## APPENDIX B: TERMINOLOGY

**Anaerobic lagoon:** Large, earthen ponds that hold the liquid fraction after mechanical or gravity-driven separation of the manure plus wash water.

**Anaerobic digester:** A system in which manure, wash water, and often other waste materials, anaerobically decompose, resulting in methane (biogas) production. This methane is collected and either flared (combusted so that it is released as climate-neutral carbon dioxide) or used to power a generator for on-farm energy (electricity, heat, or both) production. Designs range from covers for existing anaerobic lagoons to temperature-controlled batch reactors (IPCC 2006).

**Barn floors:** Heterogeneous surfaces in barns, including paved or slatted-floor areas for livestock movement, farmer access, and manure management, as well as stalls with some sort of soft bedding where livestock can rest.

**Confined animal feeding operation (CAFO):** Livestock farms that raise a large number of animals in corrals, pens, barns, or a combination of these structures.

**Corral:** A dirt-floored dry lot.

**Daily spread:** Manure is collected daily from the confinement facility and applied to pasture or cropland (IPCC 2006).

**Dairy cow:** A cow that has calved and is lactating or is briefly between lactations.

**Dairy heifer:** A cow that has not yet calved.

**Deep pit:** Manure is captured from a confinement facility, usually through a slatted floor, and stored in a tank with minimal water addition (IPCC 2006).

**Dry lot:** An open, paved or unpaved confinement area with few plants from which manure is scraped and removed periodically (IPCC 2006).

**Enteric fermentation:** Microbially mediated, anaerobic digestion of plant material in the rumen of ruminant livestock.

**Hardstandings:** Areas with solid surfaces, such as concrete, used as corrals or as temporary holding pens.

**Liquid manure:** Manure with less than 20% solid content (IPCC 2006).

**Loafing pens:** Corrals on dairies in which milk cows rest and which are typically dirt-floored and spread with some sort of bedding material, often dried manure solids.

**Manure management:** Activities related to collecting and storing manure on livestock facilities.

**Slurry tanks/settling pond:** Storage containers filled with unseparated, minimally diluted livestock manure with a higher solids content than anaerobic lagoons.

**Soil management:** Activities related to fertilization, cultivation, and organic matter addition of agricultural soils.

**Solid manure:** Manure with greater than 20% solid content (IPCC 2006).

## REFERENCES

- Adviento-Borbe, M Arlene, Eileen Fabian Wheeler, Neil E. Brown, Patrick A. Topper, R.E. Graves, V.A. Ishler, and G.A. Varga. 2010. "Ammonia and Greenhouse Gas Flux from Manure in Freestall Barn with Dairy Cows on Precision Fed Rations." *Transactions of the ASABE* 53 (4): 1251–1266.
- Aguilera, Eduardo, Luis Lassaletta, Andreas Gattinger, and Benjamin S. Gimeno. 2013a. "Managing Soil Carbon for Climate Change Mitigation and Adaptation in Mediterranean Cropping Systems: A Meta-Analysis." *Agriculture, Ecosystems & Environment* 168 (March): 25–36. doi:10.1016/j.agee.2013.02.003.
- Aguilera, Eduardo, Luis Lassaletta, Alberto Sanz-Cobena, Josette Garnier, and Antonio Vallejo. 2013b. "The Potential of Organic Fertilizers and Water Management to Reduce N<sub>2</sub>O Emissions in Mediterranean Climate Cropping Systems. A Review." *Agriculture, Ecosystems & Environment* 164 (January): 32–52. doi:10.1016/j.agee.2012.09.006.
- Ahn, H.K., W. Mulbry, J.W. White, and S.L. Kondrad. 2011. "Pile Mixing Increases Greenhouse Gas Emissions during Composting of Dairy Manure." *Bioresource Technology* 102 (3) (February): 2904–9. doi:10.1016/j.biortech.2010.10.142.
- Amon, B., V. Kryvoruchko, T. Amon, and S. Zechmeister-Boltenstern. 2006. "Methane, Nitrous Oxide and Ammonia Emissions during Storage and after Application of Dairy Cattle Slurry and Influence of Slurry Treatment." *Agriculture, Ecosystems & Environment* 112 (2–3) (February): 153–162. doi:10.1016/j.agee.2005.08.030.
- Bjorneberg, D.L., A.B. Leytem, D.T. Westermann, P.R. Griffiths, L. Shao, and M.J. Pollard. 2009. "Measurement of Atmospheric Ammonia, Methane, and Nitrous Oxide at a Concentrated Dairy Production Facility in Southern Idaho Using Open-Path FTIR Spectrometry." *Transactions of the ASABE* 52 (5): 1749–1756.
- Borhan, M.S., Sergio Capareda, Saqib Mukhtar, William B. Faulkner, Russell McGee, and Calvin B. Parnell. 2011a. "Determining Seasonal Greenhouse Gas Emissions from Ground-Level Area Sources in a Dairy Operation in Central Texas." *Journal of the Air & Waste Management Association* 61 (7) (July): 786–795. doi:10.3155/1047-3289.61.7.786.
- Borhan, Md Saidul, Sergio C. Capareda, Saqib Mukhtar, William B. Faulkner, Russell McGee, and Calvin B. Parnell. 2011b. "Greenhouse Gas Emissions from Ground Level Area Sources in Dairy and Cattle Feedyard Operations." *Atmosphere* 2 (4) (August 9): 303–329. doi:10.3390/atmos2030303.
- Brady, N.C., and R.R. Weil. 2010. *Elements of the Nature and Properties of Soils*, 3rd ed. (Upper Saddle River, NJ: Prentice Hall).
- Brown, H.A., Claudia Wagner-Riddle, and G.W. Thurtell. 2002. "Nitrous Oxide Flux from a Solid Dairy Manure Pile Measured Using a Micrometeorological Mass Balance Method." *Nutrient Cycling in Agroecosystems* 62: 53–60.
- Burger, Martin, Louise E. Jackson, Erica J. Lundquist, Dianne T. Louie, Robin L. Miller, Dennis E. Rolston, and Kate M. Scow. 2005. "Microbial Responses and Nitrous Oxide Emissions during Wetting and Drying of Organically and Conventionally Managed Soil under Tomatoes." *Biology and Fertility of Soils* 42 (2) (July 15): 109–118. doi:10.1007/s00374-005-0007-z.
- Camarillo, Mary Kay, William T. Stringfellow, Michael B. Jue, and Jeremy S. Hanlon. 2012. "Economic Sustainability of a Biomass Energy Project Located at a Dairy in California, USA." *Energy Policy* 48 (September): 790–798. doi:10.1016/j.enpol.2012.06.020.
- Capper, J.L., R.A. Cady, and D.E. Bauman. 2009. "The Environmental Impact of Dairy Production: 1944 Compared with 2007." *Journal of Animal Science* 87: 2160–2167.
- CARB (California Air Resources Board). 2011. "California Greenhouse Gas Emissions Inventory: 2000–2009."
- CDFA (California Department of Food and Agriculture). 2013. "California Agricultural Statistics Review 2012–2013." California Department of Food and Agriculture.

- CEC (California Energy Commission). 2009. "Dairy Power Production Program: Dairy Methane Digester System Program Evaluation Report." California Energy Commission, PIER Energy-Related Research Program, CEC-500-2009-009.
- Chianese, D.S., C.A. Rotz, and T.L. Richard. 2009a. "Simulation of Methane Emissions from Dairy Farms to Assess Greenhouse Gas Reduction Strategies." *Transactions of the ASABE* 52 (4): 1313–1323.
- . 2009b. "Simulation of Carbon Dioxide Emissions from Dairy Farms to Assess Greenhouse Gas Reduction Strategies." *Transactions of the ASABE* 52 (4): 1301–1312.
- . 2009c. "Simulation of Nitrous Oxide Emissions from Dairy Farms to Assess Greenhouse Gas Reduction Strategies." *Transactions of the ASABE* 52 (4): 1325–1335.
- Collins, H.P., A.K. Alva, J.D. Streubel, S.F. Fransen, C. Frear, S. Chen, C. Kruger, and D. Granatstein. 2011. "Greenhouse Gas Emissions from an Irrigated Silt Loam Soil Amended with Anaerobically Digested Dairy Manure." *Soil Science Society of America Journal* 75 (6): 2206. doi:10.2136/sssaj2010.0360.
- Conrad, Ralf. 2007. "Microbial Ecology of Methanogens and Methanotrophs." *Advances in Agronomy* 96: 1–63. doi:10.1016/S0065-2113(07)96005-8.
- Craggs, R., J. Park, and S. Heubeck. 2008. "Methane Emissions from Anaerobic Ponds on a Piggery and a Dairy Farm in New Zealand." *Australian Journal of Experimental Agriculture* 48 (2): 142. doi:10.1071/EA07255.
- Culman, Steven W., Van R. Haden, Toby Maxwell, Hannah Waterhouse, and William Horwath. 2014. *Greenhouse Gas Mitigation Opportunities in California Agriculture: Review of California Cropland Emissions and Mitigation Potential*. NI GGMOCA R 3. Durham, NC: Duke University.
- De Gryze, Steven, Juhwan Lee, Stephen Ogle, Keith Paustian, and Johan Six. 2011. "Assessing the Potential for Greenhouse Gas Mitigation in Intensively Managed Annual Cropping Systems at the Regional Scale." *Agriculture, Ecosystems and Environment* 144: 150–158. doi:10.1016/j.agee.2011.05.023.
- De Gryze, Steven, Adam Wolf, Stephen R. Kaffka, Jeff Mitchell, Dennis E. Rolston, Steven R. Temple, Juhwan Lee, and Johan Six. 2010. "Simulating Greenhouse Gas Budgets of Four California Cropping Systems under Conventional and Alternative Management." *Ecological Applications* 20 (7): 1805–1819.
- DeLonge, Marcia S., Justine J. Owen, and Whendee L. Silver. 2014. *Greenhouse Gas Mitigation Opportunities in California Agriculture: Review of California Rangeland Emissions and Mitigation Potential*. NI GGMOCA R 4. Durham, NC: Duke University.
- DeLonge, Marcia S., Rebecca Ryals, and Whendee Silver. 2013. "A Lifecycle Model to Evaluate Carbon Sequestration Potential and Greenhouse Gas Dynamics of Managed Rangelands." *Ecosystems* 16: 962–979. doi: 10/1007/s10021-013-9660-5.
- Eagle, A.J., and L.P. Olander. 2012. *Greenhouse Gas Mitigation with Agricultural Land Management Activities in the United States — A Side-by-Side Comparison of Biophysical Potential*, 1st ed. Advances in Agronomy. New York: Elsevier
- Ellis, S., J. Webb, T. Misselbrook, and D. Chadwick. 2001. "Emission of Ammonia (NH<sub>3</sub>), Nitrous Oxide (N<sub>2</sub>O) and Methane (CH<sub>4</sub>) from a Dairy Hardstanding in the UK." *Nutrient Cycling in Agroecosystems* 60: 115–122.
- ERG (Eastern Research Group). 2000. "Calculations: Percent Distribution of Manure for Waste Management Systems." Lexington, MA.
- Firestone, Mary K., and Eric A. Davidson. 1989. "Microbiological Basis of NO and N<sub>2</sub>O Production and Consumption in Soil." In *Exchange of Trace Gases between Terrestrial Ecosystems and the Atmosphere*, edited by M.O. Andreae and David S. Schimel, 7–21. New York: John Wiley & Sons.
- Gao, Zhiling, Huijun Yuan, Wenqi Ma, Xuejun Liu, and R.L. Desjardins. 2011. "Methane Emissions from a Dairy Feedlot during the Fall and Winter Seasons in Northern China." *Environmental Pollution* 159 (May): 1183–1189. doi:10.1016/j.envpol.2011.02.003.

- Gastal, F., and G. Lemaire. 2002. "N Uptake and Distribution in Crops: An Agronomical and Ecophysiological Perspective." *Journal of Experimental Botany* 53: 789–99.
- Gerber, P.J., A.N. Hristov, B. Henderson, H. Makkar, J. Oh, C. Lee, R. Meinen, F. Montes, T. Ott, J. Firkins, A. Rotz, C. Dell, A.T. Adesogan, W.Z. Yang, J.M. Tricarico, E. Kebreab, G. Waghorn, J. Dijkstra, and S. Oosting. 2013. "Technical Options for the Mitigation of Direct Methane and Nitrous Oxide Emissions from Livestock: A Review." *Animal* 7(2): 220–234.
- Gupta, Prabhat K., Arvind K. Jha, S. Koul, P. Sharma, V. Pradhan, Vandana Gupta, C. Sharma, and Nahar Singh. 2007. "Methane and Nitrous Oxide Emission from Bovine Manure Management Practices in India." *Environmental Pollution (Barking, Essex : 1987)* 146 (1) (March): 219–24. doi:10.1016/j.envpol.2006.04.039.
- Hensen, A., T.T. Groot, W.C.M. van den Bulk, C. Flechard, and A.T. Vermeulen. 2005. "Greenhouse Gas Emissions from Dairy Farms in the Netherlands MIDAIR Contribution Report."
- Hensen, A., T.T. Groot, W.C.M. van den Bulk, A.T. Vermeulen, J.E. Olesen, and K. Schelde. 2006. "Dairy Farm CH<sub>4</sub> and N<sub>2</sub>O Emissions, from One Square Metre to the Full Farm Scale." *Agriculture, Ecosystems & Environment* 112 (February): 146–152. doi:10.1016/j.agee.2005.08.014.
- Husted, Soren. 1994. "Seasonal Variation in Methane Emission from Stored Slurry and Solid Manures." *Journal of Environmental Quality* 23: 585–592.
- IPCC (Intergovernmental Panel on Climate Change). 2006. *2006 IPCC Guidelines for National Greenhouse Gas Inventories*. Prepared by the National Greenhouse Gas Inventories Programme. Edited by Simon Eggleston, Leandro Buendia, Kyoko Miwa, Todd Ngara, and Kiyoto Tanabe. Institute for Global Environmental Strategies, Japan.
- Johansen, A., H.B. Nielsen, C.M. Hansen, C. Adreasen, J. Carlsbart, H. Hauggard-Nielsen, A. Roepstorff. 2013. "Survival of Weed Seeds and Animal Parasites as Affected by Anaerobic Digestion at Meso- and Thermophilic Conditions." *Waste Management* 33(4), 807–812. doi: 10.1016/j.wasman.2012.11.001.
- Johnson, J.S., and M.G. Kern. 1993. "Conservation Tillage Impacts on National Soil and Atmospheric Carbon Levels." *Soil Science Society of America Journal* 57: 200–210.
- Kaharabata, S.K., P.H. Schuepp, and R.L. Desjardins. 1998. "Methane Emissions from Aboveground Open Manure Slurry Tanks." *Global Biogeochemical Cycles* 12 (3): 545–554.
- Kaharabata, Samuel K., Peter H. Schuepp, and Raymond L. Desjardins. 2000. "Estimating Methane Emissions from Dairy Cattle Housed in a Barn and Feedlot Using an Atmospheric Tracer." *Environmental Science & Technology* 34 (15) (August): 3296–3302. doi:10.1021/es990578c.
- Kearney, T.E., M.J. Larkin, and P.N. Levett. 1993. "The Effect of Slurry Storage and Anaerobic Digestion on Survival of Pathogenic Bacteria." *Applied Microbiology* 74(1), 86–93.
- Kellogg, Robert L., Charles H. Lander, David C. Moffitt, and Noel Gollehon. 2000. "Manure Nutrients Relative to the Capacity of Cropland and Pastureland to Assimilate Nutrients: Spatial and Temporal Trends for the United States." *Proceedings of the Water Environment Federation* 2000. doi:10.2175/193864700784994812.
- Kinsman, R., F.D. Sauer, H.A. Jackson, and M.S. Wolynetz. 1995. "Methane and Carbon Dioxide Emissions from Dairy Cows in Full Lactation Monitored over a Six-Month Period." *Journal of Dairy Science* 78 (12) (December): 2760–6. doi:10.3168/jds.S0022-0302(95)76907-7.
- Knapp, J.R., J.L. Firkins, J.M. Aldrich, R.A. Cady, A.N. Hristov, W.P. Weiss, A.D.G. Wright, and M.D. Welch. 2011. "Cow of the Future: Research Priorities for Mitigating Enteric Methane Emissions from Dairy." Innovation Center for US Dairy.
- Kong, Angela Y.Y., Johan Six, Dennis C. Bryant, R. Ford Denison, and Chris van Kessel. 2005. "The Relationship between Carbon Input, Aggregation, and Soil Organic Carbon Stabilization in Sustainable Cropping Systems." *Soil Science Society of America Journal* 69 (4): 1078. doi:10.2136/sssaj2004.0215.
- Kong, Angela Y.Y., Steven J. Fonte, Chris van Kessel, and Johan Six. 2009. "Transitioning from Standard to Minimum Tillage: Trade-Offs Between Soil Organic Matter Stabilization, Nitrous



- Oxide Emissions, and N Availability in Irrigated Cropping Systems.” *Soil and Tillage Research* 104 (2) (July): 256–262. doi:10.1016/j.still.2009.03.004.
- Leytem, April B, Robert S. Dungan, David L. Bjerneberg, and Anita C. Koehn. 2011. “Emissions of Ammonia, Methane, Carbon Dioxide, and Nitrous Oxide from Dairy Cattle Housing and Manure Management Systems.” *Journal of Environmental Quality* 40: 1383–1394. doi:10.2134/jeq2009.0515.
- Li, Changsheng, William Salas, Ruihong Zhang, Charley Krauter, Al Rotz, and Frank Mitloehner. 2012. “Manure-DNDC: a Biogeochemical Process Model for Quantifying Greenhouse Gas and Ammonia Emissions from Livestock Manure Systems.” *Nutrient Cycling in Agroecosystems* 93 (2) (May 13): 163–200. doi:10.1007/s10705-012-9507-z.
- Mangino, Joseph, Deborah Bartram, and Amy Brazy. 2002. “Development of a Methane Conversion Factor to Estimate Emissions from Animal Waste Lagoons.” In *11th International Emission Inventory Conference*. U.S. Environmental Protection Agency.
- Marik, Thomas, and Ingeborg Levin. 1996. “A New Tracer Experiment to Estimate the Methane Emissions from a Dairy Cow Shed Using Sulfur Hexafluoride (SF<sub>6</sub>).” *Global Biogeochemical Cycles* 10 (3): 413–418.
- Massé, D.I., G. Talbot, and Y. Gilbert. 2011. “On Farm Biogas Production: A Method to Reduce GHG Emissions and Develop More Sustainable Livestock Operations.” *Animal Feed Science and Technology* 166–167: 436–445.
- McGinn, S.M., and K.A. Beauchemin. 2012. “Dairy Farm Methane Emissions Using a Dispersion Model.” *Journal of Environment Quality* 41: 73–79. doi:10.2134/jeq2011.0266.
- Meyer, D., P.L. Price, H.A. Rossow, N. Silva-del-Rio, B.M. Karle, P.H. Robinson, E.J. DePeters, and J.G. Fadel. 2011. “Survey of Dairy Housing and Manure Management Practices in California.” *Journal of Dairy Science* 94 (9) (September): 4744–4750. doi:10.3168/jds.2010-3761.
- Misselbrook, T.H, J. Webb, D.R. Chadwick, S. Ellis, and B.F. Pain. 2001. “Gaseous Emissions from Outdoor Concrete Yards Used by Livestock.” *Atmospheric Environment* 35 (31) (November): 5331–5338. doi:10.1016/S1352-2310(01)00289-8.
- Mitloehner, Frank M., Huawei Sun, and John F. Karlik. 2009. “Direct Measurements Improve Estimates of Dairy Greenhouse-Gas Emissions.” *California Agriculture* 63 (2): 79–83.
- Ngwabie, N.M., K.-H. Jeppsson, S. Nimmermark, C. Swensson, and G. Gustafsson. 2009. “Multi-Location Measurements of Greenhouse Gases and Emission Rates of Methane and Ammonia from a Naturally Ventilated Barn for Dairy Cows.” *Biosystems Engineering* 103 (1) (May): 68–77. doi:10.1016/j.biosystemseng.2009.02.004.
- O’Brien, D., L. Shalloo, J. Patton, F. Buckley, C. Grainger, and M. Wallace. 2012. “Evaluation of the Effect of Accounting Method, IPCC V. LCA, on Grass-Based and Confinement Dairy Systems’ Greenhouse Gas Emissions.” *Animal* 6 (9) (September): 1512–1527. doi:10.1017/S1751731112000316.
- Osada, Takashi, Sven G. Sommer, Preben Dahl, and Hans Benny Rom. 2001. “Gaseous Emission and Changes in Nutrient Composition during Deep Litter Composting.” *Acta Agriculturae Scandinavica, Section B - Soil & Plant Science* 51 (3): 137–142.
- Parkin, Timothy B, and Thomas C. Kaspar. 2004. “Temporal Variability of Soil Carbon Dioxide Flux: Effect of Sampling Frequency on Cumulative Carbon Loss Estimation.” *Soil Science Society of America Journal* 68: 1234–1241.
- Parton, W.J., M. Hartman, D. Ojima, and D. Schimel. 1998. “DAYCENT and Its Land Surface Submodel: Description and Testing.” *Global and Planetary Change* 19, 35–48.
- Petersen, Søren O., Barbara Amon, and Andreas Gattinger. 2005. “Methane Oxidation in Slurry Storage Surface Crusts.” *Journal of Environmental Quality* 34 (2): 455–61.
- Petersen, Søren O., and Sven G. Sommer. 2011. “Ammonia and Nitrous Oxide Interactions: Roles of Manure Organic Matter Management.” *Animal Feed Science and Technology* 166–167 (June): 503–513. doi:10.1016/j.anifeedsci.2011.04.077.

- Pfost, D., C. Fulhage, C., and O. Alber. 2001. Land Application Equipment for Livestock and Poultry Manure Management EQ 383. University of Missouri Outreach & Extension. <http://extension.missouri.edu/p/EQ383>.
- Place, S.E., and F.M. Mitloehner. 2010. "Contemporary Environmental Issues: A Review of the Dairy Industry's Role in Climate Change and Air Quality and the Potential of Mitigation through Improved Production Efficiency." *Journal of Dairy Science* 93 (8) (August): 3407–3416. doi:10.3168/jds.2009-2719.
- Poudel, D.D., W.R. Horwath, J.P. Mitchell, and S.R. Temple. 2001. "Impacts of Cropping Systems on Soil Nitrogen Storage and Loss." *Agricultural Systems* 68 (3) (June): 253–268. doi:10.1016/S0308-521X(01)00011-7.
- Rotz, C.A., F. Montes, and D.S. Chianese. 2010. "The Carbon Footprint of Dairy Production Systems through Partial Life Cycle Assessment." *Journal of Dairy Science* 93 (3) (March): 1266–1282. doi:10.3168/jds.2009-2162.
- Rotz, C. Alan, Michael S. Corson, Dawn S. Chianese, Sasha D. Hafner, Rosa Jarvis, and Colette U. Coiner. 2011. "The Integrated Farm System Model Reference Manual Version 3.4." Agricultural Research Service, U.S. Department of Agriculture.
- Safley, L.M., and P.W. Westerman. 1988. "Biogas Production from Anaerobic Lagoons." *Biological Wastes* 23: 181–193.
- . 1992. "Performance of a Dairy Manure Anaerobic Lagoon." *Bioresource Technology* 42: 43–52.
- Salas, William, Changsheng Li, Frank Mitloehner, and John Pisano. 2008. "Developing and Applying Process-Based Models for Estimating Greenhouse Gas and Air Emissions from California Dairies." California Energy Commission, PIER Energy-Related Environmental Research. CEC-500-2008-093.
- Samer, M., C. Ammon, C. Loebsin, M. Fiedler, W. Berg, P. Sanftleben, and R. Brunsch. 2012. "Moisture Balance and Tracer Gas Technique for Ventilation Rates Measurement and Greenhouse Gases and Ammonia Emissions Quantification in Naturally Ventilated Buildings." *Building and Environment* 50 (April): 10–20. doi:10.1016/j.buildenv.2011.10.008.
- San Joaquin Valley Dairy Manure Technology Feasibility Assessment Panel. 2005. *An Assessment of Technologies for Management and Treatment of Dairy Manure in California's San Joaquin Valley*. <http://www.arb.ca.gov/ag/caf/dairypnl/dmtfaprprt.pdf>.
- Six, J., R.T. Conant, E.A. Paul, and K. Paustian. 2002. "Stabilization Mechanisms of Soil Organic Matter: Implications for C-Saturation of Soils." *Plant and Soil* 241: 155–176.
- Sneath, R.W., F. Beline, M.A. Hilhorst, and P. Peu. 2006. "Monitoring GHG from Manure Stores on Organic and Conventional Dairy Farms." *Agriculture, Ecosystems & Environment* 112 (2–3) (February): 122–128. doi:10.1016/j.agee.2005.08.020.
- Snell, H.G.J., F. Seipelt, and H.F.A. Van den Weghe. 2003. "Ventilation Rates and Gaseous Emissions from Naturally Ventilated Dairy Houses." *Biosystems Engineering* 86 (1) (September): 67–73. doi:10.1016/S1537-5110(03)00113-2.
- Snyder, C.S., T.W. Bruulsema, T.L. Jensen, and P.E. Fixen. 2009. "Review of Greenhouse Gas Emissions from Crop Production Systems and Fertilizer Management Effects." *Agriculture, Ecosystems & Environment* 133 (3–4) (October): 247–266. doi:10.1016/j.agee.2009.04.021.
- Sommer, S.G., S.M. McGinn, X. Hao, and F.J. Larney. 2004. "Techniques for Measuring Gas Emissions from a Composting Stockpile of Cattle Manure." *Atmospheric Environment* 38 (28) (September): 4643–4652. doi:10.1016/j.atmosenv.2004.05.014.
- Todd, R.W., N.A. Cole, K.D. Casey, R. Hagevoort, and B.W. Auvermann. 2011. "Methane Emissions from Southern High Plains Dairy Wastewater Lagoons in the Summer." *Animal Feed Science and Technology* 166–167 (June): 575–580. doi:10.1016/j.anifeedsci.2011.04.040.
- UC (University of California Division of Agriculture and Natural Resources Committee of Experts on Dairy Manure Management). 2005. "Managing Dairy Manure in the Central Valley of California."

- U.S. EPA (Environmental Protection Agency). 2011a. *Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990–2009*. Washington, D.C.
- . 2011b. *Market Opportunities for Biogas Recovery Systems at U.S. Livestock Facilities*. Washington, D.C. <http://www.epa.gov/agstar/tools/market-oppt.html>
- Van Kessel, J.S., and J.B.I. Reeves. 2002. “Nitrogen Mineralization Potential of Dairy Manures and Its Relationship to Composition.” *Biology and Fertility of Soils* 36: 118–123.
- Van Vliet, J.P. Laurens, Naveen K. Patni, and M. Abdul Matin. 2004. “GHG Emissions from Manure in a Naturally Ventilated, Freestall Dairy Barn, Comparing Sand and Sawdust Bedding.” American Society of Agricultural and Biological Engineers Paper No. 044161. St. Joseph, MI: ASABE.
- VanderZaag, A.C., C. Wagner-Riddle, K.-H. Park, and R.J. Gordon. 2011. “Methane Emissions from Stored Liquid Dairy Manure in a Cold Climate.” *Animal Feed Science and Technology* 166–167 (June): 581–589. doi:10.1016/j.anifeedsci.2011.04.041.
- Velthof, Gerard L., Peter J. Kuikman, and Oene Oenema. 2003. “Nitrous Oxide Emission from Animal Manures Applied to Soil Under Controlled Conditions.” *Biology and Fertility of Soils* 37: 221–230. doi:10.1007/s00374-003-0589-2.
- Ward, A.J., P.J. Hobbs, P.J. Holliman, and D.L. Jones. 2008. “Optimisation of the Anaerobic Digestion of Agricultural Resources.” *Bioresource Technology* 99(17): 7928–7940. doi: 10.1016/j.biortech.2008.02.044.
- Webb, J., Peter Sørensen, Gerard Velthof, Barbara Amon, Miriam Pinto, Lena Rodhe, Eva Salomon, Nicholas Hutchings, Piotr Burczyk, and Joanne Reid. 2013. “An Assessment of the Variation of Manure Nitrogen Efficiency throughout Europe and an Appraisal of Means to Increase Manure-N Efficiency.” *Advances in Agronomy* 119: 371–442. doi:10.1016/B978-0-12-407247-3.00007-X.
- Wiens, M.J., M.H. Entz, C. Wilson, and K.H. Ominski. 2008. “Energy Requirements for Transport and Surface Application of Liquid Pig Manure in Manitoba, Canada.” *Agricultural Systems* 98 (2) (September): 74–81. doi:10.1016/j.agsy.2008.03.008.
- Yamulki, Sirwan. 2006. “Effect of Straw Addition on Nitrous Oxide and Methane Emissions from Stored Farmyard Manures.” *Agriculture, Ecosystems & Environment* 112 (2–3) (February): 140–145. doi:10.1016/j.agee.2005.08.013.
- Yiridoe, E.K., R. Gordon, B.B. Brown. 2009. “Nonmarket Cobenefits and Economic Feasibility of On-farm Biogas Energy Production.” *Energy Policy* 37: 1170–1179.
- Zeeman, G. 1991. “Mesophilic and Psychrophilic Digestion of Liquid Manure.” Wageningen University.
- Zhang, G., J.S. Strøm, B. Li, H.B. Rom, S. Morsing, P. Dahl, and C. Wang. 2005. “Emission of Ammonia and Other Contaminant Gases from Naturally Ventilated Dairy Cattle Buildings.” *Biosystems Engineering* 92 (3) (November): 355–364. doi:10.1016/j.biosystemseng.2005.08.002.
- Zhao, Y., F.M. Mitloehner, L. Chai, J.C. Ramirez, K. Wang, J. Ni, C. Diehl, E.L. Cortus, T.T. Lim, B.W., Bogan, I. Kilic, A.J. Heber. 2010. *National Air Emissions Monitoring Study: Data from Two Dairy Freestall Barns in California - Site CA5B*. Final Report. Purdue University, West Lafayette, IN, July 30.

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